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Global wetland contribution to 2000–2012 atmospheric methane growth rate dynamics

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Abstract

Increasing atmospheric methane (CH₄) concentrations have contributed to approximately 20% of anthropogenic climate change. Despite the importance of CH₄ as a greenhouse gas, its atmospheric growth rate and dynamics over the past two decades, which include a stabilization period (1999–2006), followed by renewed growth starting in 2007, remain poorly understood. We provide an updated estimate of CH₄ emissions from wetlands, the largest natural global CH₄ source, for 2000–2012 using an ensemble of biogeochemical models constrained with remote sensing surface inundation and inventory-based wetland area data. Between 2000–2012, boreal wetland CH₄ emissions increased by 1.2 Tg yr⁻¹ (–0.2–3.5 Tg yr⁻¹), tropical emissions decreased by 0.9 Tg yr⁻¹ (–3.2–1.1 Tg yr⁻¹), yet globally, emissions remained unchanged at 184 ± 22 Tg yr⁻¹. Changing air temperature was responsible for increasing high-latitude emissions whereas declines in low-latitude wetland area decreased tropical emissions; both dynamics are consistent with features of predicted centennial-scale climate change impacts on wetland CH₄ emissions. Despite uncertainties in wetland area mapping, our study shows that global wetland CH₄ emissions have not contributed significantly to the period of renewed atmospheric CH₄ growth, and is consistent with findings from studies that indicate some combination of increasing fossil fuel and agriculture-related CH₄ emissions, and a decrease in the atmospheric oxidative sink.

Introduction

The increase of methane (CH_4) in the atmosphere is responsible for approximately 20% of the radiative forcing related to contemporary climate change (Ciais *et al* 2013). Since 1850, atmospheric CH_4 concentrations have risen by more than 150%, from a pre-industrial level of 700 ppb to 1834 ppb in 2015, primarily as a result of human activities that include fossil fuel extraction and agriculture practices (Kirschke *et al* 2013, Ruddiman 2013, Dlugokencky *et al* 2015, Tian *et al* 2016). In recent assessments of the global CH_4 budget, covering the period 1980 to 2009, natural wetlands were estimated to be the largest but also most uncertain source of CH_4 , emitting between 177–284 Tg $\text{CH}_4 \text{ yr}^{-1}$ using bottom-up modeling approaches and 142–208 Tg $\text{CH}_4 \text{ yr}^{-1}$ based on top-down atmospheric inversions (Kirschke *et al* 2013). Wetland emissions now represent about 30% of the total combined natural and anthropogenic sources and are projected to increase and amplify global warming (Stocker *et al* 2013). The large differences among published source estimates results from difficulties in defining wetland CH_4 producing area, uncertainties in biogeochemical modeling of anaerobic sources, oxidative sinks, and from uncertainties in atmospheric inversions (Melton *et al* 2013, Wania *et al* 2013, Patra *et al* 2016).

The atmospheric growth rate of CH_4 exhibits complex temporal variability, because (1) the gas has a short perturbation lifetime, ~ 12 years (Prather *et al* 2012), compared to longer-lived gases such as CO_2 for which emissions accumulate in the atmosphere on centennial to millennial timescales (Ciais *et al* 2013), and (2) the sources and sinks are diverse and can change rapidly over short time periods (Dlugokencky *et al* 1999, Bousquet *et al* 2006). For example, atmospheric CH_4 concentrations increased by approximately $12 \pm 6 \text{ ppb yr}^{-1}$ during the 1980s (based on observations made by the National Oceanic and Atmospheric Administration's (NOAA) Earth System Research Laboratory (ESRL) at the Mauna Loa Observatory, MLO); however in the 1990s a slowdown in growth was observed (Dlugokencky *et al* 1999), followed by a stabilization in the atmospheric growth rate of CH_4 that began in 1999 and lasted until 2006 (Dlugokencky *et al* 2009). Starting in 2007 and continuing to 2015, atmospheric CH_4 concentrations began to increase once more, at an average rate of 6.4 ppb yr^{-1} , equivalent to 17.8 Tg yr^{-1} , (Dlugokencky *et al* 2015). The drivers responsible for the CH_4 stabilization period remain unclear and may be due to changes in the concentration of atmospheric hydroxyl (OH) radicals, the main oxidative sink for methane (Heimann 2011), but with isotopic evidence also supporting either a reduced contribution from fossil fuel emissions (Aydin *et al* 2011), or reduced emissions from natural wetlands and rice cultivation (Kai *et al* 2011). In contrast, the period of renewed atmospheric CH_4 growth

shows high latitudinal variability (Nisbet *et al* 2014), and several explanations have been proposed, including a reduction in the OH sink capacity for CH_4 (Rigby *et al* 2008), increases in Arctic and tropical wetland emissions (Dlugokencky *et al* 2009), increased fossil fuel activities related to hydraulic fracking and natural gas exploitation (Jackson *et al* 2014, Rice *et al* 2016, Turner *et al* 2016), and possible changes in agriculture, in particular livestock production (Herrero *et al* 2013). Recent observational evidence suggests that the depletion of atmospheric $\delta^{13}\text{C}$ of CH_4 since 2007 supports a storyline for increasing biogenic emissions from agriculture (Schaefer *et al* 2016) rather than a decreasing thermogenic (fossil fuel related) or pyrogenic (biomass burning) emissions (Ghosh *et al* 2015).

The role of natural wetlands in the periods of stabilization (1999–2006) and renewed growth (2007–2012) has generally been overlooked in recent global CH_4 budgets because these assessments have ended too early, i.e. 1993–2004 in the WETCHIMP ecosystem model ensemble (Melton *et al* 2013), 1980–2009 in the Kirschke *et al* (2013) study, and up to 2008 in the Dlugokencky *et al* (2009) study. An additional constraint has been the limited availability and scope of temporal wetland dynamics datasets, such as the Global Inundation Extent from Multi-Satellites Observations (GIEMS), which presently only covers 1993 to 2007 (Prigent *et al* 2007, Papa *et al* 2010) and excludes seasonal or permanent wetlands where surface inundation, or flooding, is not observed. Because a key requirement for wetlands to produce CH_4 is by anaerobic soil respiration, where saturated or flooded soil conditions limit oxygen availability and thus create a suitable environment for methanogenesis, the accurate mapping of wetland area is critically important for estimating emissions. We address this issue, and problems related to comprehensively mapping wetland types (Adam *et al* 2010, Bohn *et al* 2015), by merging dynamic satellite remote sensing data of surface inundation for the 2000–2012 period (Schroeder *et al* 2015) with a static inventory of wetlands (Lehner and Doll 2004) following the same definition for natural wetlands used in Matthews and Fung (1987) and Melton *et al* (2013). These wetland definitions include both permanently and seasonally flooded soils, and include soils with either surface inundation or sub-surface saturation or both. Our definition includes only natural freshwater wetlands that are vegetated, such as peatland systems (bogs and fens) and mineral wetlands (including swamps and marshes), and attempts to avoid double counting of wetland emissions by excluding lakes, rivers, rice cultivation, saline estuaries, salt marshes, and reservoirs, which are typically accounted for as separate CH_4 fluxes in global inventories (Saunio *et al* 2016).

An additional challenge in wetland CH_4 emission modeling is that multiple pathways for CH_4 production, consumption, and release exist within

wetlands; anaerobically produced CH_4 is released to the atmosphere after being affected by a combination of processes that include oxidation by methanotrophic bacteria in the soil before diffusion to the atmosphere via plant transport structures known as ‘aerenchyma’, ebullition or through soil pores. Wetland models represent these biogeochemical and biophysical processes with varying degrees of complexity, with some modeling approaches estimating only the net flux of CH_4 as a ratio of CO_2 to CH_4 production (Christensen *et al* 1996, Kaplan 2002) and other approaches representing multiple individual processes involved in the production, consumption, and transport of CH_4 to the atmosphere (Cao *et al* 1996, Walter *et al* 2001, Riley *et al* 2011, Zürcher *et al* 2013, Grant *et al* 2015). The various formulations of model structure, parameterization, and initialization lead to relatively high uncertainties and emphasize the need for an ensemble approach in any comprehensive evaluation of the temporal dynamics and long-term evolution of global wetland CH_4 emissions. The use of multiple biogeochemical models also allows for testing hypotheses related to either the climatic sensitivity of methane emissions, versus wetland area, or substrate limitation, for example.

To extend the record of observations, and to consider process-based uncertainties, an ensemble of eleven biogeochemical models that simulate CH_4 emissions followed a common protocol (see Methods) to provide monthly integrated global wetland CH_4 emissions at 0.5° spatial resolution from 2000–2012. Global wetland area and inundation dynamics were estimated by merging (see Methods) remote sensing based observations of daily surface inundation from the Surface Water Microwave Product Series (SWAMPS; Schroeder *et al* 2015) with the static inventory of wetland area from the Global Lakes and Wetlands Database (GLWD; Lehner and Doll 2004). This approach aimed to reduce uncertainties in wetland area estimation using prognostic approaches (Melton *et al* 2013) and also addressed known issues associated with remote sensing of surface inundation where sub-surface saturation and forested wetlands are poorly detected (Bohn *et al* 2015). Methane emissions from lakes and rice paddies (Zhang *et al* 2016a), and soil consumption of atmospheric CH_4 (Curry 2007) are excluded from our estimates and included in more a recent multi-sectorial analysis of the global CH_4 budget (Saunio *et al* 2016). The overall objectives of this study were to (i) provide an estimate up to 2012 for global wetland CH_4 emissions, (ii) quantify the role of wetlands CH_4 emissions on the stabilization period (2000–2006) and the renewed growth period (2007–2012), and (iii) partition the relative role of meteorological drivers, their teleconnections, and wetland area dynamics on wetland CH_4 emissions between 2000–2012.

Methods

Wetland area dynamics

To reduce the uncertainty for wetland area dynamics resulting from predictive modeling approaches such as TOPMODEL (Gedney and Cox 2003), we combined remote sensing and inventory data to develop a monthly global wetland area dataset. Current satellite remote sensing of wetlands uses coarse-spatial resolution passive and active microwave sensors, $\sim 25 \text{ km}^2$, that observe surface water generally not obscured by vegetation (Bohn *et al* 2015, Schroeder *et al* 2015). This includes open water (e.g. lakes, rivers and ocean) as well as surface inundated wetlands comprising mainly of open plant canopies, and thus excludes exposed wetlands with no observable surface flooding as well as surface inundated wetlands beneath closed (forest) canopies. Consequently, whereas ground-based wetland inventories estimate between 8.2 and 10.1 Mkm^2 of wetlands globally (Lehner and Doll 2004), remote sensing surface water estimates of wetlands are far lower, i.e. $\sim 6.5 \text{ Mkm}^2$ excluding coastal grid regions (Schroeder *et al* 2015). To develop a comprehensive wetland dynamics dataset, we integrated the Global Lakes and Wetlands Dataset, or GLWD (Lehner and Doll 2004), with the seasonal cycle of surface water inundation from the Surface Water Microwave Product Series (SWAMPS; Schroeder *et al* 2015).

The SWAMPS dataset maps fractional surface water dynamics using remote sensing data from multiple passive and active microwave satellite missions using a 28 day screening procedure to mask snow, ice cover, and melting snow. In our analysis, SWAMPS surface inundation was derived from the Special Sensor Microwave Imager version 1 and 2 (SSM/I v1/v2), SeaWinds-on-QuickSCAT (QSCAT) from January 2000 to October 2008, and from the European Space Agency Advanced Scatterometer (ASCAT) from November 2008 to December 2012. Land-cover data from MOD12Q1 V004 (Friedl *et al* 2010) was used to exclude permanent open water (water bodies, rivers, snow/ice) and thus avoid double counting of wetlands, with an additional global FAO land mask (Zobler 1986) applied to remove coastal grid cells where brackish and salt-water wetlands were not considered as a source of methane. The monthly SWAMPS dataset (‘fw_28_swe’) was re-projected from its native 0.25° EASE grid to a geographic 0.5° rectilinear grid (WGS84) using a conservative remapping interpolation to preserve the original wetland area.

The GLWD Level 3 dataset was first reclassified to remove Classes 1–3, lakes, reservoirs and rivers, and then aggregated by summing wetland area to 0.5° from 30 arc-second resolution. GLWD is commonly used as a benchmark for various remote sensing and wetland mapping activities because it incorporates

the highest-quality country-level inventory coverage of wetlands (Peregon *et al* 2008, Fluet-Chouinard *et al* 2014, Bohn *et al* 2015). The integration of SWAMPS and GLWD took place in three phases; first the maximum annual surface water fraction at the per-pixel level ($FwMax_{x,y}$) for the 2000–2012 period was compared with GLWD ($GLWD_{max_{x,y}}$) to estimate the relative SWAMPS detection bias ($FwMaxCor_{x,y}$).

$$FwMaxCor_{x,y} = \frac{GLWD_{x,y}}{FwMax_{x,y}} \quad (1)$$

Second, $FwMax_{x,y}$ from SWAMPS was adjusted using the $FwMaxCor_{x,y}$ correction factor from equation (1) so that the maximum surface-water fraction from SWAMPS matched the GLWD estimate equation (2). For areas approximately northwest of the Hudson Bay Lowlands, the GLWD classifies the entire region as ‘lakes’, and so in cases where the merging SWAMPS-GLWD resulted in lower wetland area, the original SWAMPS surface-water values were used equation (3). Seasonal wetlands in desert systems, mapped in the GLWD, were retained in the SWAMPS-GLWD product.

$$FwMaxGLWD_{x,y} = FwMaxCor_{x,y} FwMax_{x,y} \quad (2)$$

$$FwMaxGLWD_{x,y} = \text{if } (FwMaxGLWD_{x,y} < FwMax_{x,y}, FwMax_{x,y}) \quad (3)$$

Third, the original monthly SWAMPS surface-inundation ($Fw_{x,y,m}$) was rescaled equation (4) for each year as a fraction of that same year’s (uncorrected) maximum inundation, resulting in a unique monthly scalar (0–1) for each year ($FwScalar_{x,y,m}$). Lastly, $FwMaxCor_{x,y}$ was multiplied by the annual fractional inundation cycle, $FwScalar_{x,y,m}$ (equation 5).

$$FwScalar_{x,y,m} = \frac{Fw_{x,y,m}}{FwMax_{x,y,m}} \text{ where } m = 1..12 \quad (4)$$

$$FwCor_{x,y,t} = FwMaxGLWD_{x,y} FwScalar_{x,y,t} \text{ where } t = 1..all \text{ months} \quad (5)$$

The adjusted SWAMPS-GLWD product results in a maximum wetland area of 10.5 Mkm², and in agreement with the GLWD and other studies (Fluet-Chouinard *et al* 2014), but the product also maintains the seasonal cycle and inter-annual trends of inundation mapped by SWAMPS. Key wetland areas are retained in the SWAMPS-GLWD in areas such as Amazonia, the Congo Basin, and the Western Siberian Lowlands, which in previous studies have been poorly represented (Bohn *et al* 2015).

We also conducted a sensitivity test to account for how differences in view angle between the QSCAT and ASCAT instruments might influence trends in surface inundation. SWAMPS accounted for changes in the

angle-of-incidence between sensors by applying a time-averaged normalization approach to the backscatter retrievals (Schroeder *et al* 2015), however sensor-based offsets in grid cells with low surface inundation may affect the trends. We removed low surface-inundation grid cells (defined by their maximum annual value) using a per-pixel threshold of 0.5%, 1%, 2.5% and 5%, and compared the change in methane emissions for each scenario with no filter applied.

Modeling protocol and other driver data

A common simulation protocol was followed by each of the wetland modeling teams (listed in table S1 available at stacks.iop.org/ERL/12/094013/mmedia) using standardized climate, atmospheric CO₂ and dynamic wetland area data (used to map CH₄ producing regions), and also to specify boundary conditions for model spin-up and transient runs. CRU-NCEP v4.0 was used as the meteorology data, which includes long and shortwave radiation, air pressure, specific humidity, total precipitation, air temperature, and wind speed and direction. CRU-NCEP v4.0 combines the higher spatial resolution of CRU TS3.22 (Harris *et al* 2013) with the higher temporal resolution from the NCEP Reanalysis product (Kanamitsu *et al* 2002), to produce a meteorological forcing dataset that covers years 1901–2012 at 6 hourly temporal and 0.5 degree spatial resolution, and is used as the climate driver for biogeochemical models included in the annual Global Carbon Project CO₂ budget (Le Quéré *et al* 2015). Global atmospheric CO₂ concentrations were provided at an annual resolution for 1860–2012, with data prior to 1958 from ice cores (Joos and Spahni 2008) and after 1958 from the average of NOAA measurements at Mauna Loa (MLO) and the South Pole (SPO) stations.

Models were run to equilibrium during a spin-up phase where the first thirty years of climate data, 1901–1930, were recycled with pre-industrial CO₂ concentrations ~276 ppm). Soil texture data was prescribed using model-specific global soil databases such as the Harmonized World Soils Database (FAO/IIASA/ISRIC/ISSCAS/JRC 2012) and with pedo-transfer functions (i.e. Cosby *et al* 1984) to determine water-holding capacity. Land use (i.e. agriculture or pasture) and land-cover change were not simulated, and CH₄ emissions by fire excluded from our analysis. Modeling groups used their default vegetation distributions determined by either a dynamic vegetation model or by prescribed satellite vegetation products (Poulter *et al* 2015).

Atmospheric CH₄ observations

Atmospheric observations of CH₄ were accessed from the NOAA ESRL cooperative air sampling network (Dlugokencky *et al* 1994). We carried out a comparison of wetland CH₄ emissions with atmospheric growth rate data from surface flask measurements at MLO and with the globally averaged marine surface annual mean dataset, which uses selected sites

Table 1. Wetland methane emissions in Tg CH₄ yr⁻¹ for each of the 12 TRANSCOM regions (Gurney *et al* 2003), with codes as defined in figure 1. The emissions are presented as averaged over the stabilization period (2000–2006), the increasing period (2007–2012) and for 2012. The uncertainty range is estimated as the standard deviation of the wetland CH₄ model ensemble ($n = 11$).

Region	2000–2006	2007–2012	2012
Boreal			
Boreal N America (NABo)	25.1 ± 11.3	26.1 ± 11.8	27.1 ± 12.5
Boreal Eurasia (EuBo)	11 ± 5.3	11 ± 5.2	10.7 ± 5.2
Europe (EURO)	5.7 ± 2.5	5.9 ± 2.6	6.1 ± 2.6
Temperate			
N America (NATe)	16.2 ± 5.6	16.4 ± 5.7	17.6 ± 5.9
S America (SATE)	13.4 ± 3.6	12.1 ± 3.2	11.9 ± 3.4
Eurasia (EUTE)	15.1 ± 7.1	14.8 ± 7.2	14.9 ± 7.4
Tropical			
S America (TrSA)	38.5 ± 9.3	37.4 ± 9.2	36.8 ± 9.1
Asia (TrAs)	22.5 ± 3.7	23.2 ± 3.7	23.9 ± 3.8
Africa (TrAf)	8.4 ± 1.9	8.0 ± 1.7	8.3 ± 1.8
Semi Arid			
N Africa (NAfr)	8.5 ± 3.7	8.8 ± 3.6	8.3 ± 3.3
S Africa (SAfr)	9 ± 1.9	9.2 ± 2	9.2 ± 1.8
Australia (AUST)	2.7 ± 1.5	2.7 ± 1.4	2.6 ± 1.3
Global	184 ± 21.1	183.5 ± 23.1	185.7 ± 23.2

representative of a well-mixed marine boundary layer. In addition to anthropogenic contributions, the growth rate at MLO integrates terrestrial flux processes and has been demonstrated to be useful as a representative station for diagnosing CO₂ and CH₄ exchange between the biosphere and atmosphere (Fung *et al* 1991, Dlugokencky *et al* 1995, Bousquet *et al* 2006, Wang *et al* 2014, Meng *et al* 2015). Annual mean CH₄ concentrations from 2000–2012 were first detrended, removing the long-term increase in CH₄ concentrations following 2006, because we were interested in evaluating the role of interannual climate variability on CH₄ emissions (assuming minimal variability in OH at interannual timescales) and then using a conversion of 2.78 Tg CH₄ per ppb to estimate changes in the ‘atmospheric burden’ (Fung *et al* 1991). The interannual variability of CH₄ concentrations and emissions was then calculated as $Y_i + Y_{(i+1)}$ (where $Y = \text{year}$). As could be expected, the variability in the MLO observations was more highly correlated with wetland CH₄ emission variability than with the globally averaged observations, where the averaging across multiple, mainly marine, stations across latitudes partly dampens the contribution from land to inter-annual variability. Thus, in the following, only the MLO observations are used to discuss the contribution of the modeled fluxes to atmospheric variability.

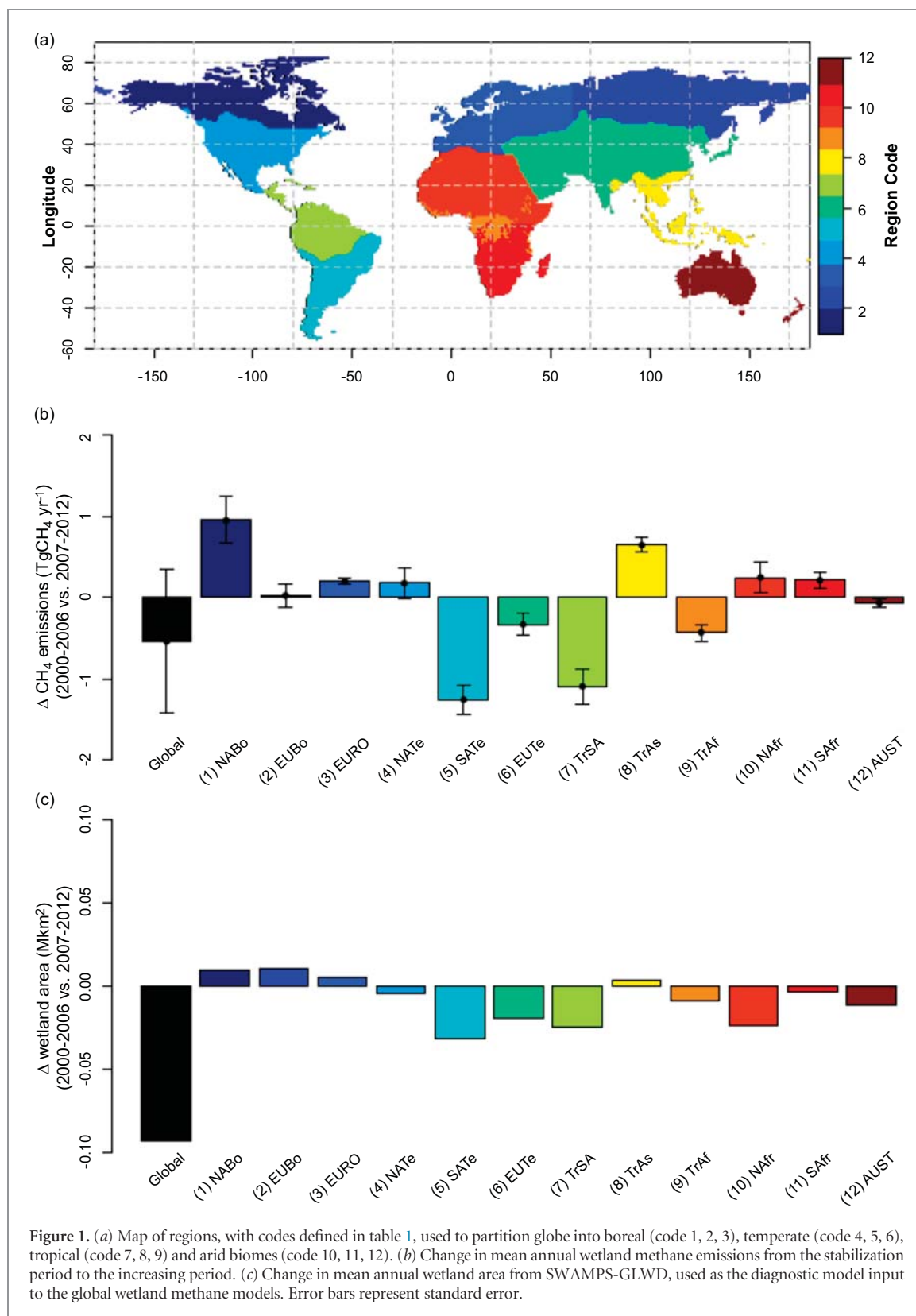
Results

Global and regional trends in wetland CH₄ emissions (2000–2012)

For the stabilization period (2000–2006), global wetland CH₄ emissions were estimated at 184 ± 21 Tg CH₄ yr⁻¹, where the uncertainty is estimated as one standard deviation of the model ensemble mean. Wetland emissions remained statistically similar during the period of renewed growth (2007–2012) at 183 ± 23 Tg CH₄ yr⁻¹, with a slightly larger value in the last year of analysis, 2012, of 186 ± 23 Tg CH₄ yr⁻¹

(table 1). For both time periods, tropical biomes, defined in figure 1(a) as regions 7 to 9, dominated the global flux with representative 2012 emissions, for example, of 69 ± 12 Tg CH₄ yr⁻¹, followed by boreal (44 ± 19 Tg CH₄ yr⁻¹), temperate (44 ± 10 Tg CH₄ yr⁻¹), and sub-tropical biomes (20 ± 5 Tg CH₄ yr⁻¹). The global fluxes are consistent with a range of previously published estimates using satellite based approaches, i.e. 170 Tg CH₄ yr⁻¹ (Bloom *et al* 2010), atmospheric inversions, i.e. 149–159 Tg CH₄ yr⁻¹ (Ghosh *et al* 2015) and 165 ± 9 Tg CH₄ yr⁻¹ (Bousquet *et al* 2011), process-based models, i.e. 190 ± 39 Tg CH₄ yr⁻¹ (Melton *et al* 2013) and a combination of inversion and process-model, i.e. 172 Tg CH₄ yr⁻¹ (Spahni *et al* 2011).

Between the two time periods (the 2000–2006 ‘stabilization’ and the 2007–2012 ‘renewed growth’ periods), no statistically significant change in the average model ensemble emissions (two-sided Student’s t -test; $\alpha = 0.1$) was found at the global scale or regionally (figure 2(a) and figure 3(a)). Among individual models, the change in global emissions between 2000–2006 and 2007–2012 ranged from a 5.4 Tg CH₄ yr⁻¹ decrease for ORCHIDEE to an increase of 4.8 Tg CH₄ yr⁻¹ for LPJ-MPI, with an ensemble average change of -0.5 ± 0.9 Tg CH₄ yr⁻¹ (figure 1(b)). At the regional scale (figure 1(b)), an increase in boreal wetland CH₄ emissions of 1.2 ± 0.3 Tg CH₄ yr⁻¹ was found for the ensemble, with only CLM4.5 estimating reduced emissions of 0.2 Tg CH₄ yr⁻¹ and with LPJ-MPI providing the largest increase of 3.5 Tg CH₄ yr⁻¹. Six of the individual models had statistically significant increasing trends for boreal CH₄ emissions (linear regression, $p < 0.1$) and of these six, all models agreed with an increase in emissions occurring for June–August (JJA) and September–November (SON). A decrease in tropical emissions of 0.9 ± 0.3 Tg CH₄ yr⁻¹ between 2000–2006 and 2007–2012 was observed across the model ensemble, with just two models estimating an increase



(CLM4.5 and LPJ-MPI, 0.1 and 1.1 $\text{Tg CH}_4 \text{ yr}^{-1}$, respectively) and ORCHIDEE estimating the largest decrease of 3.1 $\text{Tg CH}_4 \text{ yr}^{-1}$. Four of the individual models had a statistically significant ($p < 0.1$) decrease in tropical CH_4 emissions between JJA and SON. Changes in tropical wetland emissions were sensitive to the filtering of the low-surface inundated wetlands,

carried out to detect inter-sensor bias, but no statistically significant change was detected (table S3). A decrease in temperate regional emissions of $1.4 \pm 0.4 \text{ Tg CH}_4 \text{ yr}^{-1}$ and almost no change in semi-arid emissions ($0.4 \pm 0.2 \text{ Tg CH}_4 \text{ yr}^{-1}$) was also obtained for the ensemble, but with a larger spread across models than for the boreal and tropical regions.

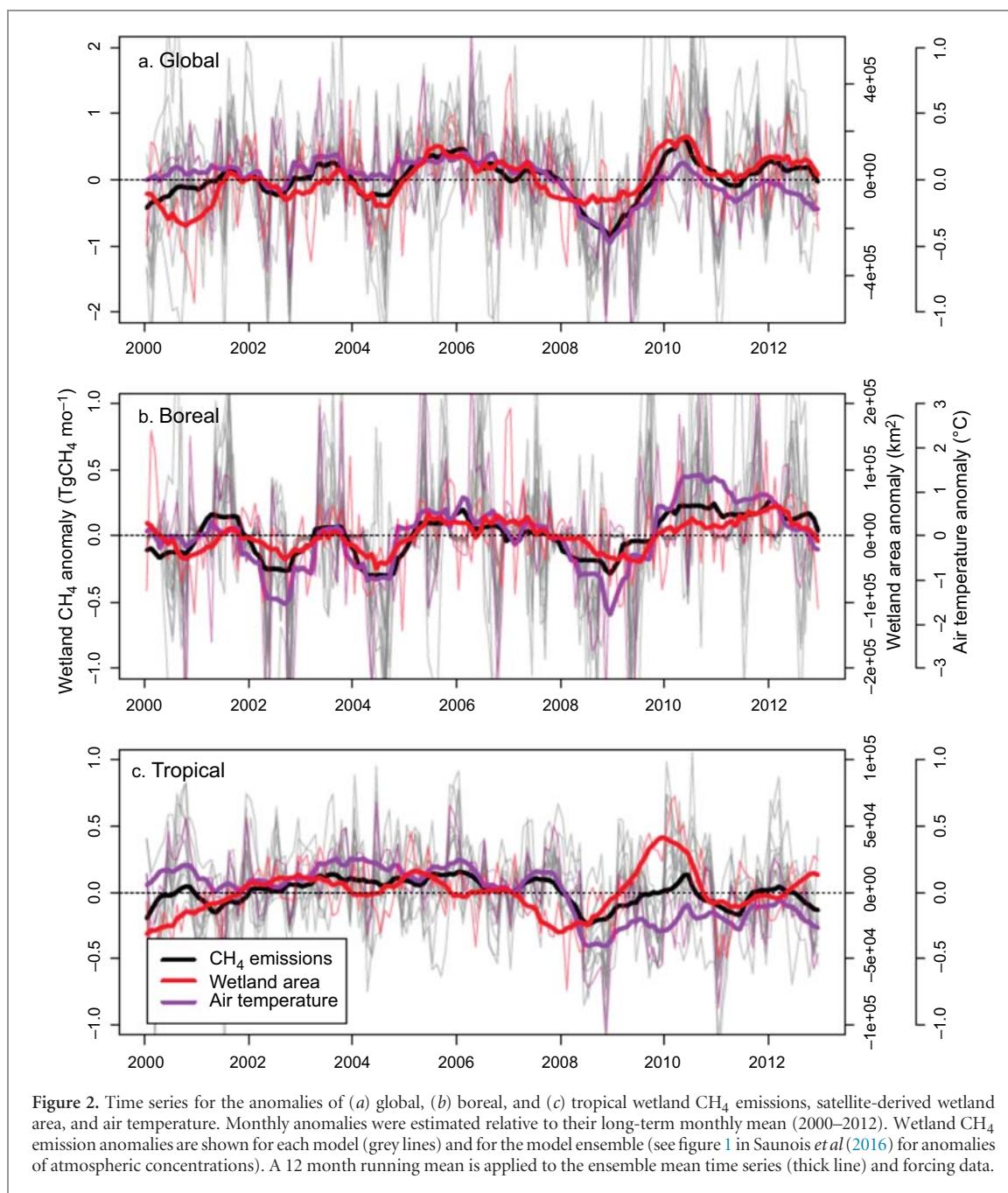


Figure 2. Time series for the anomalies of (a) global, (b) boreal, and (c) tropical wetland CH₄ emissions, satellite-derived wetland area, and air temperature. Monthly anomalies were estimated relative to their long-term monthly mean (2000–2012). Wetland CH₄ emission anomalies are shown for each model (grey lines) and for the model ensemble (see figure 1 in Saunio *et al* (2016) for anomalies of atmospheric concentrations). A 12 month running mean is applied to the ensemble mean time series and forcing data.

Climatic and physical drivers of regional CH₄ trends

During the 2000–2012 period, a linear regression analysis with climate forcing based on the Climate Research Unit, CRU TS3.22 (Harris *et al* 2013), masked to match wetland containing grid cells only (as an average over the 2000–2012 time period), revealed variable spatial and seasonal trends in precipitation and air temperature (for annual trends, see figure 3(c) and (d)). Total global December–February (DJF) precipitation increased by 2.5 mm yr⁻¹ ($p < 0.05$) but did not change significantly in other seasons (figure 3(c)). Increasing boreal winter (DJF) precipitation contributed to about half of the annual global increase, 1.4 mm yr⁻¹ ($p = 0.1$), with other boreal seasons showing no change, tropical DJF precipitation increased by 8.2 mm yr⁻¹ ($p = 0.01$), and semi-arid DJF precipitation increased

by 1.8 mm yr⁻¹ ($p = 0.04$). Global annual air temperature over wetlands was nearly constant (figure 3(d)), and increased at a rate of 0.02 °C yr⁻¹ in JJA ($p < 0.05$) and 0.04 °C yr⁻¹ in SON ($p < 0.05$). The change in global air temperature was mainly due to increasing air temperature in boreal regions with a significant ($p < 0.1$) rate of increase of 0.06 °C yr⁻¹ from 2000–2012 in SON, and in tropical biomes where air temperature also increased slightly in JJA and SON at a rate of 0.02 °C yr⁻¹ ($p < 0.1$). Cloud cover (not shown) increased by 0.1% yr⁻¹ ($p = 0.05$) between March–May (MAM) in the tropics and decreased during boreal MAM by -0.2% yr⁻¹ ($p = 0.03$).

Average annual maximum global wetland area for the merged SWAMPS-GLWD was 10.5 million km² (see Methods for comparison with original SWAMPS)

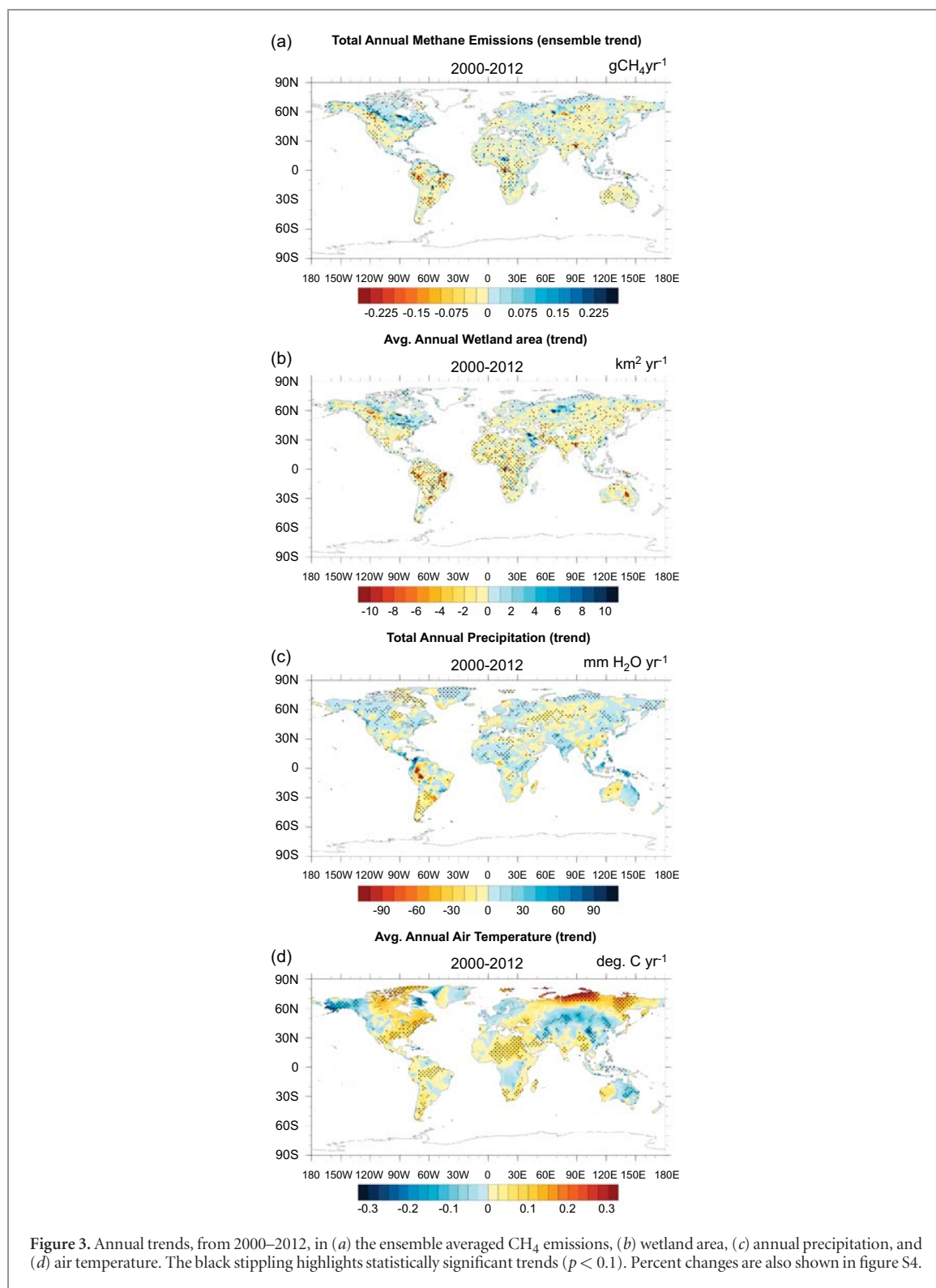


Figure 3. Annual trends, from 2000–2012, in (a) the ensemble averaged CH_4 emissions, (b) wetland area, (c) annual precipitation, and (d) air temperature. The black stippling highlights statistically significant trends ($p < 0.1$). Percent changes are also shown in figure S4.

and was in agreement with the GLWD inventoried global wetland area (Lehner and Doll 2004) used as the basis for several benchmarking activities (Fluet-Chouinard *et al* 2014, Bohn *et al* 2015). Globally and regionally, the SWAMPS-GLWD dataset had a similar seasonal phase (figure S1) for wetland area as GIEMS ($R^2 > 0.85$ for all except the semi-arid region) yet SWAMPS-GLWD had larger seasonal amplitude because of the addition of permanent wetlands from

GLWD. The overlapping time period for GIEMS and SWAMPS, years 2000–2007, showed no significant trends globally or regionally for both datasets. Between the CH_4 stabilization and renewed growth periods, global mean annual wetland area decreased by 93 000 km^2 (2% of average annual wetland area, figure 1(c) and figure 3(b)). At the seasonal scale, a large part of the decrease was explained by negative JJA trends in wetland area, where a statistically

significant decrease of $27\,000\text{ km}^2\text{ yr}^{-1}$ was observed ($p < 0.01$). Much of the seasonal decrease was explained by statistically significant changes in tropical, temperate, and semi-arid JJA wetland area of -4600 , -1400 , and $-6000\text{ km}^2\text{ yr}^{-1}$, respectively, with additional changes in DJF tropical ($-3500\text{ km}^2\text{ yr}^{-1}$) and semi-arid ($-7300\text{ km}^2\text{ yr}^{-1}$) wetland area observed. In the boreal regions, wetland area increased by 3000 and $16\,400\text{ km}^2\text{ yr}^{-1}$ in DJF and SON, respectively ($p < 0.01$). Overall, a complex pattern of regional and seasonal contributions in declining global wetland area was observed, consistent with decadal and multi-decadal observations of land-water storage and open-water bodies (Dieng *et al* 2015, Donchyts *et al* 2016), and with tropical wetland area decreasing (-3.4%) and boreal wetland increasing (1.8%) in area (figures 2(b) and (c)) between the stabilization and the renewed growth periods.

Sensitivity of CH₄ emissions to climate and wetland area

A partial correlation analysis was carried out to determine the effect of wetland area and climate on CH₄ emissions, and to determine the interaction between local climate and large-scale climatic teleconnections, including the Multivariate El Niño Index (MEI) and the North Atlantic Oscillation (NAO), on regional wetland area dynamics. Partial correlation analysis is an appropriate statistic to provide estimates on the correlation coefficient for a set of variables while simultaneously controlling for their interactions. The resulting partial correlations, r , range from -1 to 1 with absolute values closer to unity reflecting higher explanatory power, either with a negative or positive relationship between the independent and dependent variables. Monthly time series for each variable were correlated for the period 2000–2012 and the data were not detrended beforehand because there were no significant trends detected.

The MEI and NAO represent two major global climatic teleconnections, with the MEI linking Pacific sea surface temperature anomalies (lagged by one month) with a warming and drying in tropical regions in its positive El Niño phase and a wetting of mid-latitude arid regions in its negative La Niña phase (Wolter and Timlin 1993). The MEI is similar in its temporal dynamics to the Oceanic Niño Index that uses sea surface temperature anomalies from the Niño 3.4 region. In contrast, the NAO measures the difference in air pressure between the Icelandic low and Azores high (Barnston and Livezey 1987), reflecting mid-to-high latitude climates, with a positive NAO characterized by above average annual temperature and wet winters in Eastern North America and northern Europe and below-average temperatures in the arctic. In contrast, during the negative NAO phase, cooler and drier than average conditions persist in eastern North America and northern Europe, with warmer than average conditions in the Arctic.

For the model ensemble, variability in global CH₄ emissions was most highly correlated with wetland area ($r = 0.64$), followed by temperature ($r = 0.37$) and with negligible correlations for precipitation ($r = 0.09$) and cloud cover ($r = 0.11$). A two to three month lag between the CH₄ emissions response and climate increased the precipitation correlation by a small amount, from 0.09 (with no time lag) to 0.11 with a one month lag. Monthly to seasonal scale lags have also been observed in atmospheric inversion and hydrologic studies (Papa *et al* 2015, Ribeiro *et al* 2016, Wilson *et al* 2016) where transit time of water within a basin and other hydrologic processes, such as evapotranspiration, decouple the more immediate interactions between precipitation and emissions. At the regional scale, wetland area was also the most important variable for CH₄ emissions, with a correlation of 0.89 and 0.72 in tropical and temperate regions, respectively. In contrast, for the individual models, global CH₄ emission for JULES and LPJ-MPI was more highly correlated with air temperature than with wetland area due to their greater temperature sensitivity than other models, whereas the remaining models were correlated first with wetland area, and then with air temperature followed by smaller precipitation or cloud cover correlations. The ranking of global wetland area as the main driver of CH₄ production, followed by temperature and then precipitation was similar for the boreal, tropical, temperate, and arid regions, and consistent with results from a multi-model CH₄ sensitivity experiment carried out by Melton *et al* (2013). Overall, the higher air temperature sensitivity of CH₄ emissions was responsible for moderate correlations, with varying time lags (t minus number months, n), with the MEI for boreal ($r_{t-0} = -0.16$), tropical ($r_{t-6} = 0.33$), and temperate emissions ($r_{t-3} = 0.24$), with the NAO also only weakly correlated with the model ensemble for boreal regions ($r_{t-3} = -0.13$). At the global scale, the MEI and NAO were most highly correlated with CH₄ emissions with a five-month lag, $r_{t-5} = 0.26$ and $r_{t-5} = 0.08$, respectively, slightly lower than in previously published studies (Bousquet *et al* 2006, Hodson *et al* 2011).

Wetland area dynamics from the SWAMPS-GLWD dataset at global and regional scales were moderately correlated with air temperature (masked for wetland grid cells), $r_{t-0} = 0.24$ globally, and $r_{t-0} = 0.33$ for boreal regions, suggesting surface-flooding increased following seasonal permafrost thaw under warmer temperatures (Schuur *et al* 2015). Precipitation (also masked for wetland grid cells) was weakly correlated with global wetland area ($r_{t-0} = -0.11$), and with wetland area in temperate ($r_{t-0} = 0.18$), tropical ($r_{t-0} = -0.12$), and arid regions ($r_{t-0} = -0.15$). The introduction of time lags (up to $+6$ months) in the climate variables did not significantly improve the correlations with wetland area except in semi-arid regions where a one month lag increased the precipitation correlation with wetland area ($r_{t-1} = 0.30$). These results highlight the importance of incorporating sub-grid cell

topographic variation, as well as cell-to-cell interactions, when modeling feedbacks between hydrologic flow paths and surface inundation dynamics. At the global scale, the MEI was positively, albeit weakly, correlated with wetland area ($r_{t-5} = 0.33$) mainly because of the relationship with tropical ($r_{t-5} = 0.29$) and temperate wetland dynamics ($r_{t-5} = 0.25$). Boreal regions were negatively correlated with MEI with a one-month lag ($r_{t-1} = -0.14$) and both the MEI and NAO positively correlated with wetland area in temperate regions ($r_{t-0} = 0.27$ and $r_{t-0} = 0.15$, respectively). The moderate correlations, compared with earlier studies, were partly due to the short time series, where between 2000–2012, the NAO was mainly in negative phase (meaning below-average precipitation in mid-high latitudes, cooler eastern North America and northern European temperatures, and warmer arctic conditions), and no large El Niño events, whereas a record magnitude La Niña lasted from late 2009 to 2011 (Evans and Boyer-Souchet 2012). In addition, our regional definitions may also interfere with the strength of the teleconnection correlations by introducing a mix of biome types with varying climatic responses (Zhang *et al* 2015).

Discussion

High latitude increases and low latitude decreases in CH₄ emissions

From 2000–2012, global wetland emissions appear to have remained stable and with regional increasing and decreasing trends closely compensating for one another, with no net contribution to the observed renewed atmospheric growth rate. By shifting the period of comparison to 2003–2005 and 2010–2012 to evaluate the sensitivity of our definition for the stabilization and renewed growth periods, we find only a slightly larger increase in emissions, from 185 ± 22 Tg CH₄ yr⁻¹ to 186 ± 24 Tg CH₄ yr⁻¹, an average 1.23 ± 1.1 Tg CH₄ increase and also not large enough to explain the 2007 renewed atmospheric CH₄ growth rate of ~ 17 Tg CH₄ yr⁻¹. Additionally, the increase between the 2003–2005 and 2010–2012 periods is not robust and almost entirely driven by just one model, which also has the highest temperature sensitivity, LPJ-MPI (12.6 Tg CH₄ yr⁻¹ increase). The increase in boreal emissions from 2000–2012 appears to be closely linked to both increasing air temperature and wetland area, with an anomalously warm event in 2007 (Bruhwiler *et al* 2014). In high latitude regions, evidence for warming air temperature is well documented and the feedbacks between increasing air temperature, sea-ice cover loss, and terrestrial CH₄ emissions is becoming increasingly clear (Karl *et al* 2015, Parmentier *et al* 2015). Finer-temporal and spatial remote-sensing based analyses are also consistent with the evidence presented here for a net increase in boreal wetland area and CH₄ emissions from 2003 to 2011 (Watts *et al* 2014). Overall, these changes are

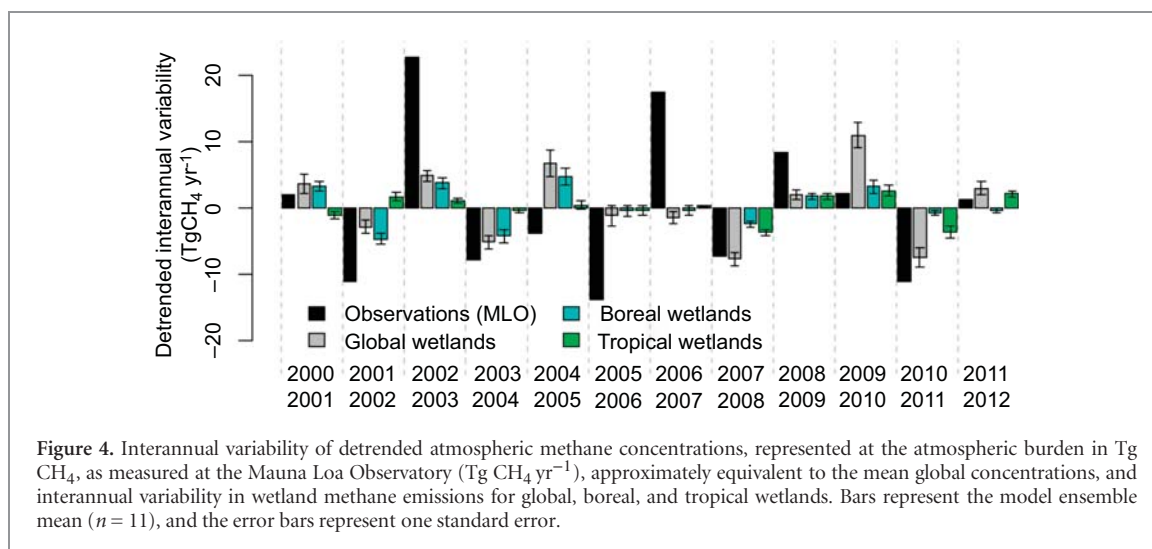
consistent with field observations (Sweeney *et al* 2016) and with what could be expected from projected climate change and warming impacts on high latitude systems that link temperature sensitivity as a dominant control on arctic wetland CH₄ emissions (Schaefer *et al* 2011, Chen *et al* 2015, Schuur *et al* 2015).

In tropical regions, high interannual variability in precipitation makes detecting decadal scale carbon-cycle trends challenging (Jung *et al* 2010, Zhang *et al* 2015). In terms of wetland area dynamics, Papa *et al* (2010) reported a decrease of $19\,600$ km² yr⁻¹ in tropical surface inundation between 1993 and 2005 based on the GIEMS data, and losses of tropical surface inundation appear to have continued through 2012 at a rate of 4000 km² yr⁻¹ (Schroeder *et al* 2015). As a consequence, declining tropical wetland CH₄ emissions have been found in a range of studies using GIEMS, for example, Meng *et al* (2015), who found a decline of 1.68 Tg CH₄ yr⁻¹ from 1993–2004. However, even with models that used fixed or static, i.e. the GLWD, rather than dynamic areal extent of wetlands, declining tropical wetland CH₄ emissions were simulated (Zhu *et al* 2015), suggesting that trends in climatic drivers that force changes in wetland area may be an equally important constraint on tropical CH₄ production. Over Amazonia and the Congo Basins, large consecutive droughts, in 2005 and 2010, combined with regional warming, have resulted in widespread declines in tropical forest canopy greenness (Hilker *et al* 2014, Zhou *et al* 2014). These Amazonian droughts are superimposed on an intensification of the hydrologic cycle in the wet season (Gloor *et al* 2013) rather than an increase in the duration of the wet season throughout the year. Declining precipitation trends between 2010–2012 were observed in Western Amazonia and Eastern Congo, but over Southeast Asia, increases in precipitation were observed (figure 3(c)). Degradation of global wetlands due to human activities is also a large component in declining wetland function (Petrescu *et al* 2015, Donchyts *et al* 2016), and losses of tropical wetland area due to drainage are the highest globally, ranging up to 2% per year (Davidson 2014, Papa *et al* 2015).

Outside of tropical regions, declining wetland area in temperate regions also coincided with long-term drying of soils from 1950–2005 (Mueller and Zhang 2015), however the precipitation trends from 2000–2012 of relevance in this study suggest soil moisture actually increased in this time period (figure 3(c)). Semi-arid regions, i.e. eastern Australia and South America, showed decreasing annual precipitation trends, despite large swings in seasonal precipitation related to a series of strong La Niña events (Boening *et al* 2012).

Role of teleconnections on interannual variability of wetland area and CH₄ emissions

ENSO has previously been highlighted as a key driver of interannual variability in global wetland CH₄ production (Bousquet *et al* 2006, Hodson *et al* 2011,



Dalsøren *et al* 2015). Here, we find a possibly lower role for ENSO in driving global wetland CH₄ production that is due in part to i) the duration of the brief time series where no strong El Niño was observed and ii) the use of a new integrated wetland–surface–water dataset. Previous analyses have used longer time series, such as Hodson *et al* (2011), who scaled modeled soil-moisture to wetland area based on a GIEMS calibrated hydrologic model. For the time period 1950–2005 they found a slightly higher global correlation of global CH₄ emissions with ENSO, $R^2 = 0.39$ (with a three-month lag) and $R^2 = 0.56$ for the tropics. In addition, using an atmospheric inversion, Bousquet *et al* (2006) partitioned CH₄ emissions to anthropogenic and natural sources for the period 1984–2003. The study concluded that the dominant role of CH₄ surface sources was from high interannual variability of wetland area that was synchronized with ENSO. The longer 1950–2005 time period includes a wider range of both positive and negative phase ENSO events, whereas the 2000–2012 period evaluated here includes two major La Niña and only moderate El Niño events.

In addition, many studies have relied on GIEMS surface inundation data to constrain wetland areal dynamics, and have found GIEMS to be highly correlated with ENSO (Prigent *et al* 2007). While the new SWAMPS-GLWD dataset used here was found to enhance seasonal variation in wetland area, the dataset also partially decoupled the interannual surface-water variability from climate because of the integration of permanently inundated wetlands (with no surface flooding) from GLWD, to address known limitations in microwave remote sensing of wetlands, particularly in forested areas (Bohn *et al* 2015). We evaluated how the development of the SWAMPS-GLWD wetland dynamics dataset affected interannual variability (IAV) of wetland CH₄ emissions by comparing with the detrended observations of atmospheric CH₄ variability from MLO. We found the observed IAV at MLO to range from -13 to 22 Tg CH₄ yr⁻¹ from 2000 to 2012 (figure 4). In comparison, the IAV of

the SWAMPS-GLWD driven wetland model ensemble ranged from -13 to 19 Tg CH₄ yr⁻¹, and across models, the range varied from small IAV (-8 to 1 Tg CH₄ yr⁻¹ for CTEM) to large IAV (3 to 19 Tg CH₄ yr⁻¹ for ORCHIDEE). Compared to observations, the sensitivity of the model ensemble results provide confidence in the use of SWAMPS-GLWD for partially driving a CH₄ IAV consistent with previous top-down and isotopic studies, e.g. Bousquet *et al* (2006), that demonstrate wetland CH₄ emissions explain a large portion of the IAV in atmospheric growth ($r = 0.46$ for the model ensemble, with the individual models ranging from $r = 0.2$ (TRIPLEX-GHG) to $r = 0.6$ (SDGVM)). Notably, the contribution of boreal wetlands to global CH₄ IAV appears to decline from 2000–2012 relative to an increase from tropical contributions (figure 4), however, overall we found a trend toward decreasing IAV in the observed CH₄ growth rate. Wildfires, not considered in this study, can contribute between 10 – 20 Tg CH₄ yr⁻¹ of emission IAV, however no significant trend over time has been observed to date (van der Werf *et al* 2006, van der Werf *et al* 2010, Worden *et al* 2013). Additionally, year-to-year variability in the atmospheric oxidative sink for methane may also affect variability in growth rate anomalies (Rigby *et al* 2008).

Uncertainties from additional biogenic CH₄ sources

The depletion of atmospheric $\delta^{13}\text{C}$ of CH₄ since 2007 presents three scenarios, (i) a change in average biogenic wetland $\delta^{13}\text{C}$ source signature, (ii) an overall increasing biogenic source, (iii) a decreasing thermogenic or pyrogenic source, or some combination of all (Kirschke *et al* 2013). Thermogenic, or fossil-fuel related emissions, are unlikely to have decreased in recent years (Bergamaschi *et al* 2013, Nisbet *et al* 2014), and recent studies based on isotopic $\delta^{13}\text{C}$ confirm a large biogenic source (Dlugokencky *et al* 2011, Rice *et al* 2016, Schaefer *et al* 2016). In addition to the wetland types considered in this study, there are several additional sources of biogenic emissions that

could contribute to the depletion of atmospheric $\delta^{13}\text{C}$. These include river systems (Bastviken *et al* 2011, Borges *et al* 2015), lakes (Verpoorter *et al* 2014, Tan and Zhuang 2015), and agriculture (Leff *et al* 2004, Chen *et al* 2013). For example, taken together, river and lake system CH_4 emissions are highly uncertain and are estimated to emit as much as 100 ± 50 Tg $\text{CH}_4 \text{ yr}^{-1}$ (Bastviken *et al* 2011), or the equivalent of $\sim 30\%$ – 50% of global wetland emissions, and would require a reassessment of other source terms to close the global methane budget (Saunio *et al* 2016). These emission hotspots are also geographically distributed across arctic (Walter Anthony *et al* 2014), temperate (Chen *et al* 2013) and tropical systems (Borges *et al* 2015). The temporal response of agricultural CH_4 emissions (excluding biomass burning) is poorly understood, yet agriculture accounts for about 30% of total wetland CH_4 emissions (Kirschke *et al* 2013) and is produced from rice cultivation and enteric fermentation of livestock ruminants. These agricultural emissions can change on annual to decadal time scales in response to climate (Li *et al* 2002), but also in response to farming practices where land management can rapidly respond to socio-economic drivers (Chen *et al* 2013) and contribute to atmospheric IAV and to long-term trends (Tian *et al* 2016).

Reducing biogenic source uncertainty

By using a multi-model approach to investigate the temporal trends and spatial patterns in global CH_4 emissions, the model uncertainty can be quantified more robustly. Here, the sources of uncertainty can be partitioned to (i) driver data, (ii) model structure, and (iii) parameter uncertainty. By providing a consistent set of climate, atmospheric CO_2 , and wetland area data, the model spread was reduced from 123 Tg $\text{CH}_4 \text{ yr}^{-1}$ (WETCHIMP, which used a similar model ensemble) to 80 Tg $\text{CH}_4 \text{ yr}^{-1}$ (this study). This reduction highlights that uncertainties in wetland area are almost equally important to our mechanistic understanding of *in situ* CH_4 production and consumption processes. Combining SWAMPS and GLWD led to wetland area estimates consistent with more detailed regional estimates for Amazonia (Wilson *et al* 2007, Draper *et al* 2014, Hess *et al* 2015), southeast Asia (Hooijer *et al* 2010), and high-latitude systems, such as the Western Siberia Lowlands (Bohn *et al* 2015, Zhang *et al* 2016b).

Model structure is another key source of uncertainty (Wania *et al* 2013, Xu *et al* 2016a), as illustrated by the range of temperature-emission sensitivities for the current model ensemble. About half of the models used here (JULES, LPJ-wsl, ORCHIDEE, SDGVM, CTEM) were based on the semi-empirical model approach of Christensen *et al* (1996), whereas the other models (LPX-Bern, LPJ-MPI, CLM4.5, TRIPLEX-GHG) were based on more mechanistic first order approaches based

on the framework developed by Walter *et al* (2001), see table S2 for a summary. One topic of large uncertainty are the oxidative processes that consume CH_4 (Ridgwell *et al* 1999), and that may change over time and alter the CO_2 : CH_4 production ratios used in the semi-empirical approaches (Curry 2007). However, there was no clustering of model structure in terms of global or regional emission trends.

In addition to meteorological and wetland area interannual variability, atmospheric CO_2 rose by 24 ppm to 393 ppm from 2000 to 2012 (based on observations from Mauna Loa, MLO). Net primary production in carbon cycle models tends to respond positively to trends in elevated CO_2 (Hickler *et al* 2008), and would be expected to provide a sustained increase in substrate in the form of soil organic carbon for anaerobic processes to produce CH_4 . A strong CO_2 driven response in CH_4 emissions was not observed by the ensemble mean because of the high IAV of climate and wetland area that appear to be more limiting for CH_4 emissions than substrate. Over longer timescales, i.e. multi-decadal to centennial, a strong CO_2 feedback on CH_4 emissions is expected, with simulated increases in global emissions of up to $73\% \pm 49\%$ at 857 ppm CO_2 (Melton *et al* 2013).

Lastly, model parameters are difficult to robustly estimate because CH_4 production occurs in complex landscapes where anaerobic soil conditions can be very heterogeneous. To estimate emissions at scales of 50 km^2 or larger, where CH_4 production may be occurring in small topographic depressions, remains a large challenge (Lara *et al* 2015, Shi *et al* 2015). While the area-weighted monthly-average flux estimates for the model ensemble ranged within observations, i.e. from 2.7 to 3.9 $\text{g CH}_4 \text{ m}^{-2} \text{ month}^{-1}$ globally, 0.7 to 3.1 $\text{g CH}_4 \text{ m}^{-2} \text{ month}^{-1}$ for boreal wetlands (observations 2–4 $\text{g CH}_4 \text{ m}^{-2} \text{ month}^{-1}$ in boreal systems (Turetsky *et al* 2014), and 5.2 to 8.2 $\text{g CH}_4 \text{ m}^{-2} \text{ month}^{-1}$ for tropical wetlands (observations 0.1 to 29 $\text{g CH}_4 \text{ m}^{-2} \text{ month}^{-1}$ (Sjogersten *et al* 2014). Benchmarking of process models with flux tower measurements or airborne campaigns remains critical for improving model structure and parameters (Miller *et al* 2016) and addressing scaling artifacts that may obscure non-linear methane production and consumption processes.

Conclusions

Key findings

Interpreting the interannual and decadal dynamics of the CH_4 atmospheric growth rate has presented significant challenges over the past three decades, with the sources and sinks remaining poorly understood (Kirschke *et al* 2013). Using an ensemble of global wetland models constrained with satellite and inventory based surface inundation and wetland area seasonality and trends, we now provide a comprehensive and updated estimate of the role of wetlands in the recent

increase of the atmospheric growth rate that began in 2007. We show that the role of wetlands in the renewed period of atmospheric CH₄ growth appears minimal to non-existent, and that:

- At the global scale, wetland CH₄ emissions have remained constant from 2000–2012 at 184 ± 22 Tg CH₄ yr⁻¹ but that significant spatial variability in trends are masked by the global perspective (figure S2 and S3).
- In boreal regions, increasing CH₄ emissions corresponds to increasing wetland area and air temperature, whereas in the tropics, decreasing wetland area and large variability in precipitation has led to decreased emissions.
- At global and the regional scales defined in our study, the role of climatic teleconnections such as ENSO and the NAO are smaller than what has been reported in previous work; however, we confirm that the IAV of the atmospheric growth rate is largely explained by wetlands.
- The interannual variability in global wetland emissions is dominated by boreal regions from 2000–2006 and then with increasing contribution from tropical regions possibly coinciding with larger droughts over Amazonia and the Congo Basin (figure 4). However, there has been no consistent shift in the IAV of wetland CH₄ emissions over the 2000–2012 time period (figure 4).
- The range of the modelled interannual variability in global wetland emissions in 2007–2012 is similar to the IAV observed at the MLO station, while it is less than observed for 2000–2006. Therefore, the period 2000–2006 is anomalous not only due to the absent trend in the growth rate of atmospheric CH₄ concentrations, but also due to anomalously high IAV not fully explained by natural wetland emissions.
- Our results, interpreted in the context of a depletion in atmospheric $\delta^{13}\text{C}$ observed since 2007, suggests that either a shift in $\delta^{13}\text{C}$ biogenic source signature occurred or other agricultural biogenic sources are required to explain the recent and sustained atmospheric increase in CH₄, or that, less likely, a decrease in thermogenic and pyrogenic emission has occurred. This is consistent with recent work of Schaefer *et al* (2016) who present isotopic evidence suggesting an increasing role of livestock and agriculture in the growth rate of atmospheric CH₄.
- The pattern of increasing high-latitude emissions and possibly decreasing to stable tropical emissions are consistent with climate change projections that forecast a general increase in boreal air temperatures and a decrease in tropical precipitation (Scholze *et al* 2006). Thus the past decade presents an observational test case for climate and socio economic impact studies on CH₄ production (Lawrence *et al* 2015, Petrescu *et al* 2015).
- To reduce uncertainties in wetland dynamics mapping we recommend that (1) multi-platform remote sensing using both radar and optical observations are integrated at higher spatial resolution to resolve issues associated with low-detection probabilities in closed-forest canopy regions, (2) that inter-sensor calibration and effects on inter-annual and seasonal trends are clearly accounted for, and (3) that ground-based wetland inventories are continually updated and made available to benchmark and calibrate remote sensing algorithms, and with clear terminology to avoid double counting.

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