

1 The origin and future of an endangered crater lake endemic;
2 phylogeography and ecology of *Oreochromis hunteri* and its invasive
3 relatives

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12

13 **Abstract**

14 Cichlids of the genus *Oreochromis* (“Tilapias”) are intensively used in aquaculture around the
15 world. In many cases when “Tilapia” were introduced for economic reasons to catchments
16 that were home to other, often endemic, *Oreochromis* species, the loss of native species
17 followed. *Oreochromis hunteri* is an endemic species of Crater Lake Chala on the slopes of
18 Mount Kilimanjaro, and is part of a small species flock in the upper Pangani drainage system
19 of Tanzania. We identified three native and three invasive *Oreochromis* species in the region.
20 Reconstructing their phylogeography we found that *O. hunteri* is closely related to, but distinct
21 from the other members of the upper Pangani flock. However, we found a second, genetically
22 and phenotypically distinct *Oreochromis* species in Lake Chala whose origin we cannot fully
23 resolve. Our ecological and ecomorphological investigations revealed that the endemic *O.*
24 *hunteri* is currently rare in the lake, outnumbered by each of three invasive cichlid species. It
25 is mitochondrially, phenotypically and trophically distinct from all others. The occurrence of
26 the formerly abundant *O. hunteri* in such small numbers, its narrow habitat restriction and its
27 limited morphological variability suggest recent population decline and loss of niche breadth
28 in this critically endangered endemic cichlid species.

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32 Introduction

33 Biodiversity provides a multitude of ecosystem services (Watson et al. 2005) and for sustainable
34 development, it is therefore essential to conserve biodiversity. The cichlids in the East African Lakes
35 provide a biodiversity hotspot in freshwater ecosystems with several hundred endemic species in each
36 of the big lakes Tanganyika, Malawi and Victoria (Genner et al. 2004) and many more endemic species
37 in numerous smaller lakes (Wagner et al. 2014). While most of the species richness consists of the
38 smaller haplochromine cichlids, most of the interest in fisheries and aquaculture is with the larger
39 oreochromine cichlids, often referred to as “Tilapia”. The oreochromine cichlids nonetheless made
40 several small radiations, especially in some of the smaller lakes (Ford et al. 2015, Schliewen et al. 1994,
41 Trewavas et al. 1972) that do not have haplochromines (Seehausen 2007). They also produced several
42 endemic, often allopatric species in Eastern African rivers and small lakes (Trewavas 1983).

43 In tropical Africa, fish is one of the most important protein sources and in Eastern Africa millions of
44 people depend heavily on fishery-related activities for their daily food or income (DeGraaf & Garibaldi
45 2014, FAO 2007, Ogutu-Ohwayo et al. 1997). As most East African aquatic ecosystems are dominated
46 by cichlid fish, they provide important economic and food resources. Since the 1950s, the optimization
47 of fishery yields led to the introduction of Nile Tilapia (*Oreochromis niloticus*, Linnaeus, 1758) in many
48 places of Africa outside the species’ distribution range, with negative impacts on the native
49 *Oreochromis* species reported in many cases through competition and hybridization (Deines et al.
50 2014, Genner et al. 2013). *Oreochromis niloticus* is one of the world’s most important species for food
51 provisioning from aquaculture. It has been distributed for aquaculture not only within Africa, but
52 worldwide (Casal 2006, Deines et al. 2016) and has established numerous wild populations in tropical
53 and subtropical latitudes, including in China (Shuai et al. 2015), in the USA (Lowe et al. 2012), in India
54 (Singh et al. 2013) and in Brazil (de Azevedo-Santos et al. 2011) to name only a few. Detrimental
55 impacts on the local ecosystems due to resource competition with the native fish or habitat alterations
56 due to the grazing behaviour, have been reported from various countries (Crutchfield 1995, Daga et
57 al. 2016, Singh et al. 2013, Zhang et al. 2017). If introduced into catchments that have their own native

58 *Oreochromis* species, *O. niloticus* causes threats additionally through hybridization with the local
59 species (Angienda et al. 2011, D'Amato et al. 2007, Deines et al. 2014, Firmat et al. 2013).

60 In this study we investigate the status of a little known endemic crater lake species, *Oreochromis*
61 *hunteri* (Günther, 1889) of Lake Chala. We investigate aspects of its origins and distinctiveness within
62 the endemic species flock of the upper Pangani drainage, its distinctiveness versus several invasive
63 cichlids, and its ecology in the face of the presence of other cichlid species that have invaded the lake
64 in recent decades. *Oreochromis hunteri* is thought to be part of a small species flock of three or four
65 taxa in the upper Pangani drainage system. Lake Chala is a small (4.2 km²) but deep (90m) crater lake
66 with habitats of rock and lava sand. It is positioned at the eastern flank of Mount Kilimanjaro (3.317°S,
67 37.699°E) at 877 m above sea level, shared between Kenya and Tanzania. Inflow and outlet are both
68 subterranean and the water in the lake is very clear (Secchi depth ~6.6 m). Four different cichlid species
69 occur in Lake Chala: the endemic *Oreochromis hunteri*, an undescribed *Oreochromis* species
70 (sometimes erroneously referred to as *O. korogwe* (Lowe, 1955), here referred to as *O. sp.* “blue
71 head”), the widely distributed eastern red-breast tilapia *Coptodon rendalli* (Boulenger, 1897) and a
72 haplochromine cichlid of the riverine *Astatotilapia bloyeti* (Sauvage, 1883) complex. The latter three
73 most likely were introduced between 1951 and 1977 (Dadzie et al. 1988, Lowe 1955). Due to the co-
74 occurrence of an endemic cichlid species that had several thousand years to explore and exploit the
75 different ecological niches within Lake Chala, and one closely and two more distantly related species
76 that only arrived about 40 – 70 years ago, Lake Chala provides a system for investigating impacts of
77 introduced species on endemic cichlid communities of isolated lakes.

78 We reconstructed the phylogeography of the native and the recent colonizations, and investigated
79 ecology and ecomorphology of all four cichlid species in Lake Chala, as well as the oreochromine
80 cichlids of the upper Pangani/Ruvu drainage. We wanted to know if and how niches are partitioned
81 between the endemic and the invasive species to make predictions for the future of this unique little
82 ecosystem and its endemic cichlid species. We experimentally fished all habitats and several depth

83 gradients in Lake Chala, and combined studying habitat distribution of cichlids, analysis of 14
84 morphological traits and $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ stable isotopes.

85

86 Material and Methods

87 Fish sampling

88 We sampled *Oreochromis* from four locations in North-Eastern Tanzania, in Lake Chala, in the Nyumba
89 ya Mungu Reservoir (dammed Pangani River), in the Ruvu River (the main inflow into the Nyumba ya
90 Mungu Reservoir) and in Lake Babati in November 2014 (figure 1). In Lake Chala gill nets with mesh
91 sizes from 12-22 mm were set at various depths in the benthic habitat in Lake Chala. Mesh sizes of the
92 nets were chosen to catch fish of comparable size of all species present in Lake Chala. This might have
93 led to a slight bias towards smaller individuals in the larger *Oreochromis* and *Coptodon* species. Nets
94 set below 47 m and nets set in the limnetic habitat (4 times three joined nets with 16, 19 and 22mm
95 mesh size) remained empty. Fish were carefully removed from the net and photographed from their
96 left side in a custom designed photo cuvette with a colour reference bar, a scale bar and a standardized
97 grey background. The fish was afterwards euthanized with an overdose of phenoxyethanol diluted in
98 lake water. From the right side of each fish, we removed the pectoral fin and fixed it in pure analytical
99 ethanol (100%) for DNA extractions, and we removed a piece of the epaxial muscle dorsal of the lateral
100 line, removed the skin, and dried the muscle tissue in an oven (60°C for 24 to 48 hours) for stable
101 isotope analysis. Whole fish were subsequently fixed in 4% formalin (buffered with borax) and were
102 later rinsed with water and stepwise transferred to 75% ethanol. Additionally, we bought several larger
103 individuals of *O. hunteri* from a local fisherman on the lake, who was fishing with hook and line. With
104 the same methods (gill netting) we caught and processed the fish from Lake Babati. In the Ruvu River,
105 the fish were caught with hook and line, and the fish from the Nyumba ya Mungu were bought from
106 the local fish market. Like this, we obtained samples from 4 sites and taxonomically identified 3 native
107 and 3 invasive *Oreochromis* species.

108

109 DNA extraction, sequencing and phylogenetic reconstruction

110 DNA was extracted from fin clips of *Oreochromis* from Lake Chala (18 *O. sp.* “blue head”, 13 *O. hunteri*),
111 Nyumba ya Mungu (6 *O. jipe* (Lowe, 1955), 2 *O. esculentus* (Graham 1928), 1 *O. niloticus*), the Ruvu
112 River (14 *O. sp.* “blue head”, 3 *O. pangani* (Lowe, 1955), 1 *O. niloticus*) and Lake Babati (2 *O. niloticus*,
113 4 *O. sp.*), and *Coptodon rendalli* from Lake Chala (N=4) and Nyumba ya Mungu (N=2). We used the
114 Qiagen DNeasy Blood&Tissue-Kit and sequenced 830 bp of the mitochondrial D-loop region.

115 We aligned our mitochondrial sequences in BioEdit v7.2.5 together with mitochondrial D-loop
116 sequences of most other *Oreochromis* species occurring in Eastern Tanzania from GenBank (D'Amato
117 et al. 2007, Kinaro et al. 2016, Nagl et al. 2001, Nyingi et al. 2009). A Templeton, Crandall and Sing (TCS)
118 network was built and visualized in PopART (Bandelt et al. 1999), to investigate and visualize the
119 relationship among the different populations and species.

120

121 Morphology and trophic ecology

122 For 27 *Oreochromis hunteri*, 74 *O. sp.* “blue head” and 10 *Coptodon rendalli* from Lake Chala, 20 *O.*
123 *jipe*, 5 *O. esculentus*, 5 *O. niloticus*, 2 *C. rendalli* and 10 *Astatotilapia* from the Nyumba ya Mungu
124 Reservoir, 6 *O. pangani*, 16 *O. sp.* “blue head” and 1 *O. niloticus* from the Ruvu River, we measured 15
125 linear morphometric distances that have proven powerful for quantifying taxonomically and
126 ecologically relevant morphometric variation in cichlids (Barel et al 1977, Witte 1984) using digital
127 callipers: standard length (SL), head length (HL), head width (HW), body depth (BD), lower jaw length
128 (LJL) and width (LJW), snout length (SnL), snout width (SnW), cheek depth (ChD), preorbital depth
129 (POD) and width (POW), interorbital width (IOW), eye length (EyL) and depth (EyD), and premaxillary
130 pedicel length (PPL). The same measurements for 274 *Astatotilapia* from Lake Chala were obtained
131 from Moser et al. (2018). All measurements were log-transformed and size-corrected by using the
132 standardized residuals of the pooled linear regression of each log-transformed trait against log-SL. The
133 largest *O. hunteri* that we bought from the fisherman at Lake Chala (SL > 19cm) were excluded as they

134 were four to eleven cm larger than the next smaller fish. All analyses, if not stated differently, were
135 performed in R 3.2.1 (R-Core-Team 2015).

136 To quantify morphological variation and differentiation, we performed principal component analyses.
137 To prevent that the PCA axes are mainly driven by the taxa with larger sample sizes, we used only ten
138 individuals of the species where sample sizes exceeded ten and all individuals for all other species for
139 the calculation of PCA axes. All other individuals were subsequently projected into the morphospace.
140 As an estimate of morphological differentiation, we calculated Bhattacharyya distances between the
141 different species in the PC1-PC2-morphospace and tested for significance with Hotelling's T-squared
142 test. To test for differences in morphological variation, we calculated morphospace occupation for
143 each species. For this purpose, we subsampled 1000 times ten individuals per species and calculated
144 the morphospace occupation using the R-function `chull`. We report the mean of these 1000
145 permutations and used the permutations to assess whether morphospace occupation is significantly
146 different between species. To investigate which traits are most distinct between *O. hunteri*, *O. sp.*
147 "blue head", *O. jipe* and *O. pangani*, we additionally performed pairwise linear discriminant analyses
148 (LDA).

149 For all fish from Lake Chala, we performed stable isotope analyses. Dried muscle tissue was ground
150 into fine powder using a QIAGEN Tissuelyzer II with five-millimetre stainless steel beads (QIAGEN). Of
151 this powder, 0.25 - 0.35 mg was used to analyse $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values in a ThermoFisher Scientific Flash
152 2000 elemental analyser coupled through a ConFlo IV interphase to a Delta V Advantage Isotope Ratio
153 Mass Spectrometer (IRMS). To quantify ecological variation and differentiation, we calculated
154 Bhattacharyya distances between the different species based on the stable isotope ratios and
155 tested for significance with Hotelling's T-squared test.

156

157 Results

158 Taxonomy, diversity and phylogeography of *Oreochromis* in Lake Chala and the nearby
159 Ruvu/Pangani system

160 We found evidence for three native (*Oreochromis hunteri*, *O. jipe* and *O. pangani*) and three non-native
161 (*O. niloticus*, *O. esculentus* and *O. sp.* "blue head") oreochromine species in the Pangani/Ruvu drainage,
162 including Lake Chala (figure 2). The D-loop haplotype network revealed that *O. hunteri* is most closely
163 related to the geographically nearby Pangani-system endemics, *O. jipe* and *O. pangani*, but does not
164 share any haplotypes with either of these (figure 3). This confirms earlier morphology-based
165 suggestions, that these three species form a "Pangani system flock" of closely related species
166 (Trewavas 1983). We did not encounter any *Oreochromis hunteri* outside Lake Chala, neither did we
167 encounter either of the other native Ruvu/Pangani taxa in Lake Chala.

168 However, we found a second, genetically distinct *Oreochromis* species in Lake Chala whose origin we
169 could not fully resolve. This species, which we refer to as *O. sp.* "blue head", does not belong to the
170 Ruvu/Pangani radiation in its mitochondrial lineage but is more closely related to *O. urolepis* from the
171 Tanzanian East Coast rivers (figure 3) and also shared the distinctive large midlateral blotches with the
172 latter (figure 2). In *O. sp.* "blue head" we found two very distinct haplotype groups that differ from
173 each other by 12 mutations (in 347 bp), suggesting that the population may be a mix of introductions
174 from two distinct source populations. We identified one individual from Lake Chala phenotypically as
175 *O. niloticus* or an *O. niloticus*-hybrid but its mitochondrial haplotype was one of the two *O. sp.* "blue
176 head" haplotypes. We also found two non-oreochromine cichlids in Lake Chala, *Coptodon rendalli* and
177 *Astatotilapia cf. bloyeti*. Both also occur in the Ruvu/Pangani system, where *Astatotilapia cf. bloyeti* is
178 native and *C. rendalli* has been introduced.

179 In the Ruvu River we found matching phenotypic and genetic evidence for the presence of three
180 *Oreochromis* species: the native *O. pangani*, the introduced *O. niloticus* and the unidentified *O. urolepis*
181 clade taxon that we also found in Lake Chala (*O. sp.* "blue head", both haplotypes). In the Nyumba ya
182 Mungu Reservoir finally, we found matching phenotypic and genetic evidence for presence of three
183 *Oreochromis* species too, but here these were *Oreochromis jipe* (supposedly introduced from Lake Jipe
184 further upstream in the catchment) and introduced *O. niloticus* and *O. esculentus*.

185 In Lake Babati we found phenotypic and genetic evidence for the presence of *O. niloticus*. Further we
186 found mitochondrial haplotypes belonging to *O. spilurus* and *O. esculentus* (figure 3), but specimens of
187 these two haplogroups could not be distinguished from each other phenotypically.

188

189 Ecological and phenotypic species differentiation and niche partitioning

190 We find that in Lake Chala each of the invasive *O. sp* "blue head", *C. rendalli* and *A. cf. bloyeti* are more
191 abundant than the native *O. hunteri* (figure 4). With the exception of a few fish caught shallower, *O.*
192 *hunteri* seems to be restricted to a depth range between 20 and 45m, i.e. the deeper half of the
193 oxygenated habitat. Even at those depths, it was outnumbered between 3 and 20 times by other cichlid
194 species (figure 4). However, our stable isotope data suggest significant feeding niche partitioning
195 between *O. hunteri* and all other species (figure 5). In fact, all four species in Lake Chala differ
196 significantly from each other in their stable isotope signatures (figure 5, table 1). *O. hunteri* feeds on
197 slightly more depleted carbon sources than *O. sp.* "blue head" and *C. rendalli* but fully overlapping with
198 *A. cf. bloyeti*. On the other hand, *O. hunteri* feeds at a slightly higher trophic level than *O. sp.* "blue
199 head" and *C. rendalli* and lower than *A. cf. bloyeti*. The latter species feeds predominantly on
200 invertebrates (insect larvae, zooplankton) and small fish (Moser et al. 2018) and its mean $\delta^{15}\text{N}$ value is
201 10.6. *Coptodon rendalli* is known as a macrophyte eater elsewhere but probably grazes mostly on dead
202 wood and organic debris in Lake Chala where macrophytes are nearly absent. Its mean $\delta^{15}\text{N}$ -value in
203 Lake Chala is about 8.0, slightly higher than *O. sp.* 'blue head' with a mean of 7.7, and nearly one trophic
204 level below *Astatotilapia*, as expected if it feeds mainly on plant matter. The mean $\delta^{15}\text{N}$ -value of *O.*
205 *hunteri* in our data is about 8.4, with the largest fish slightly but not much higher. This suggests that *O.*
206 *hunteri*, despite no evidence of a unique habitat, may have a unique feeding niche that includes both
207 vegetable and animal matter.

208 Our morphological analysis reveals that *O. hunteri* is clearly distinct eco-morphologically from all other
209 species in the system too. Within Lake Chala all four species are morphologically distinct from each
210 other (table 1), but the strongest overlap in morphospace is between *Oreochromis hunteri* and the

211 invasive *Oreochromis* sp. “blue head” (figure 6a). The main difference between these species are a
212 longer head, a less deep body and a narrower snout in *O. hunteri* compared to *O. sp.* “blue head” (table
213 2). A corresponding pattern is detected outside Lake Chala, where all species except *Oreochromis jipe*
214 and *O. pangani* can be morphologically distinguished in the PCA (table 3, figure 6b) with the highest
215 PC1 and PC2 values for the native *Oreochromis* species (*O. pangani* and *O. jipe*). Whereas PC1 (-ChD
216 and -LJL) mainly separates *Oreochromis* (small cheek depth and short lower jaws) from *Astatotilapia*
217 and *Coptodon*, PC2 (-IOW, -POD, -SnW, -BD) separates the Pangani *Oreochromis* (*O. hunteri*, *O. jipe*
218 and *O. pangani* with comparably narrow interorbital width and snout width, and low body depth) from
219 the introduced *Oreochromis* and *Coptodon* (comparably wide interorbital width and snout width, and
220 deep body depth) from *Astatotilapia*. Together, these two axes explain 74.5% of the overall variation
221 (45.7% and 28.7% respectively).

222 Within the endemic Pangani species flock, *O. hunteri* is morphologically distinct from both *O. jipe* and
223 *O. pangani* (with a longer and wider head, shorter lower jaws and smaller eyes in *O. hunteri*; table 2),
224 whereas we could not detect significant differentiation between the latter two (table 4, figure 6c). We
225 notice that whereas in the Pangani River (Nyumba ya Mungu Reservoir), the native species (*O. jipe*)
226 occupies significantly more of the morphospace than the newly arrived *O.* “blue head”, the latter
227 occupies more (though not significantly more) of the morphospace than *O. hunteri* in Lake Chala (table
228 5, figure 6a and 6b).

229

230 Discussion

231 *Oreochromis hunteri* and the oreochromine cichlids of the Pangani/Ruvu drainage

232 *Oreochromis hunteri* Günther 1889 is one of the least known African cichlids, despite it being the type
233 species of the genus that contains the world’s most important aquaculture fish. When Lowe (1955)
234 described a small flock of *Oreochromis* species endemic to the upper Pangani River drainage of
235 Tanzania, she was not sure if *O. hunteri* belonged to this species flock. Trewavas (1983) suggested,
236 based on morphological grounds, that *O. hunteri* belonged to the Pangani flock and that it was closely

237 related to *O. jipe*. Trewavas characterized the members of the flock by high numbers of vertebrae,
238 scales in the lateral line and dorsal fin rays, and suggested that four species belonged to it: *O. pangani*
239 of rivers Ruvu and Pangani, *O. girigan* of Lake Jipe, *O. jipe* of Lake Jipe and possibly the Ruvu/Pangani
240 River, and *O. hunteri* of Lake Chala. This little flock had not been studied since, and never in a
241 phylogeographic context until recently (Dieleman et al. 2018)

242 We collected three endemic and three invasive *Oreochromis* species in the upper Pangani drainage
243 including Lake Chala. We find strong support for the proposed monophyly of the upper Pangani flock
244 in the mitochondrial lineage (figure 3), confirming the results of Dieleman et al. (2018), who sequenced
245 two mitochondrial markers. Their proximity in morphospace further supports their close relationship
246 (figure 6c). Among the three endemic species that we obtained, *O. hunteri* is genetically and
247 phenotypically distinct and indeed restricted to Lake Chala. *O. jipe* appears to be common in the
248 Nyumba ya Mungu Reservoir. This was already reported by Trewavas (1983) but it has never been
249 completely clear whether its original range comprised of the Pangani/Ruvu River and Lake Jipe or was
250 confined to Lake Jipe (Lowe 1955, Trewavas 1983). Finally, we found *O. pangani* in the Ruvu River, the
251 main inlet to the reservoir, but we did not find *O. jipe* there. We cannot rule out that some *O. pangani*
252 were among our samples from Nyumba ya Mungu, but all our larger individuals had 4 anal fin spines,
253 a defining trait of this species according to Lowe (1955) and Trewavas (1983), and the typical coloration
254 of *O. jipe*, whereas all our Ruvu River samples had 3 anal spines (Lowe 1955, Trewavas 1983).

255 Beside the introduced *O. niloticus* that we found in the Ruvu River and the reservoir, and *O. esculentus*
256 which we found in the reservoir and that were already reported in the Ruvu/Pangani by Lowe (1955),
257 we found a third most likely non-native oreochromine cichlid species in the system that we refer to as
258 *O. sp. "blue head"*. *O. sp. "blue head"* occurs in two very divergent mitochondrial haplotypes. Both
259 were found in Lake Chala as well as in the Ruvu River. One of these two haplotypes suggests, identity
260 with or a close relationship to *O. urolepis urolepis* (figure 3). As the second haplotype is divergent by
261 twelve mutations and not closely related to any known *Oreochromis* species, the origin and the

262 evolutionary relationship of *O. sp.* "blue head" remain partly unresolved but it is likely that it emerged
263 from admixture of two divergent lineages in the *O. urolepis* clade.

264

265 The origin of *Oreochromis hunteri*

266 *O. hunteri* appears to have evolved a more limnetic life style when compared to typical *Oreochromis*
267 including its close allopatric relatives *O. jipe* and *O. pangani*, but also the distant relative *O. sp.* "blue
268 head" with which it is now sympatric in Lake Chala. This is apparent both from its body shape and
269 stable isotopic signatures. *O. hunteri* evolved in isolation in the crater lake for at least 19'000 years
270 (Dieleman et al. 2015). Cichlids are known for undergoing sympatric speciation into distinct ecomorphs
271 with different habitat and/or feeding adaptations. This has involved divergence into limnetic and
272 benthic forms in several oreochromine cichlids, and divergence into shallow and deep benthic forms
273 in several haplochromine cichlids. There is no evidence of such divergence within the contemporary
274 population of *O. hunteri*. Analyses of fossil teeth from Lake Chala sediment cores, and comparison with
275 contemporary teeth revealed that *O. hunteri* and *O. sp.* "blue head" together show higher
276 morphological disparity in tooth shape than the fossil assemblage. However, the fossil assemblage
277 occupies a larger volume in morphospace than the contemporary *O. hunteri* alone, with some fossil
278 teeth overlapping in morphospace with *O. sp.* "blue head" (Dieleman et al. 2015). This might indicate
279 that *O. hunteri* was ecologically more diverse historically than it is now. Especially more robust teeth
280 that may facilitate scraping of benthic algae are absent in the contemporary *O. hunteri* but dominate
281 the introduced species. To conclude whether *O. hunteri* has been more diverse in the past or whether
282 there was perhaps a second more littoral native species, more fossil data would be required.

283

284 The future of *Oreochromis hunteri*

285 Our data suggest major risks for the persistence of *O. hunteri* arising from the recent introductions of
286 four other cichlid species to Lake Chala. While currently very rare in the lake, and perhaps only present
287 as hybrids, the potential presence of *O. niloticus* is worrying given its invasiveness across large parts of

288 the world and its known impact on native *Oreochromis* elsewhere through genetic introgression
289 coupled with competitive advantages (Angienda et al. 2011, Deines et al. 2014). More immediate
290 impacts on *O. hunteri* are likely to come from the presence of the three other introduced cichlid
291 species, all of which are more abundant in Lake Chala now than is *O. hunteri*. Our data suggest massive
292 habitat overlap with all three species, and dietary overlap mainly with *O. sp.* "blue head". The latter
293 species is probably able to hybridize with *O. hunteri* (as probably is *O. niloticus*) whereas *C. rendalli* and
294 *A. cf. blozeti* most likely are too distantly related to *O. hunteri* to hybridize. Our stable isotope data,
295 our and recently published (Dieleman et al. 2018) morphological data and published tooth shape data
296 (Dieleman et al. 2015) all suggest that *O. hunteri* is phenotypically and ecologically distinct from *O. sp.*
297 "blue head". Although we had a few fish in our data that we could not clearly identify, we did not
298 detect mismatches between phenotype and mitochondrial haplotype in the individuals that we
299 identified with certainty. However, the extent of introgression and associated threats to the
300 persistence of *O. hunteri* need now to be analysed using microsatellite DNA or next generation
301 sequencing.

302 The complementarity in stable isotope signatures between *O. hunteri* and the invasive species in Lake
303 Chala is remarkable. We do not know if the current feeding and habitat niche of *O. hunteri* is its original
304 range, or if its niche has shrunk or shifted in response to the arrival of the invasive species. However,
305 several observations point to possible recent change:

306 First, in 1951 Lowe observed shoals of young feeding on algae and debris from the bottom between
307 the rocks near the shore (Lowe 1955). We have not seen any *O. hunteri* in the shallow waters despite
308 many diving hours, but we encountered all three invasive species very abundantly. In 1995, one of us
309 (OS) photographed a pelagic shoal of subadult *O. hunteri* several meters above the ground (figure 7).

310 We have not seen any such shoals in 2014.

311 Second, the *Oreochromis* teeth in the fossil assemblage studied by Dieleman et al (2015) occupied a
312 larger volume in morphospace than the contemporary *O. hunteri* teeth, with some fossil teeth
313 overlapping in morphospace with the invasive *O. sp.* 'blue head'.

314 Third, the relative morphospace occupation between the introduced and the native species is much
315 less favourable to the native *Oreochromis hunteri* in Lake Chala than for the native species *O. jipe* in
316 the Nyumba ya Mungu Reservoir. Given that *O. hunteri* had several thousand years of evolution in Lake
317 Chala (Dieleman et al. 2015), and therefore the possibility to adapt to the different niches within the
318 lake without competing with any other fish species, we would have expected *O. hunteri* to be
319 phenotypically more variable than the invasive species. This is indeed what we see in *O. jipe* compared
320 to *O. sp.* “blue head”. However, we see the opposite in Lake Chala where the invasive *O. sp.* “blue
321 head” is phenotypically significantly more diverse than *O. hunteri*. This again might indicate, that *O. sp.*
322 “blue head” displaced *O. hunteri* from parts of its morphospace and likely also its ecological niche.

323

324 Conclusions

325 We provide evidence for the presence of three endemic and three invasive *Oreochromis* species in the
326 upper Pangani drainage (Nyumba ya Mungu Reservoir, Ruvu River and Lake Chala). The three native
327 species that we collected in the Pangani/Ruvu and in Lake Chala have retained their phenotypic and
328 mitochondrial distinctiveness despite their current sympatry with several non-native species that
329 arrived since the 1950s in the Ruvu/Pangani and since the 1970s in Lake Chala. However, there is
330 serious concern for the future persistence at least of the endemic *O. hunteri* which is currently the
331 least abundant cichlid species in Lake Chala, phenotypically less diverse than its allopatric sister species
332 and the introduced sympatric species in Lake Chala, and missing from habitats where it was historically
333 recorded. Nuclear genetic marker studies are required to investigate potential introgressive
334 hybridization with the numerically dominant invasive species.

335

336

337 **Ethics.** This research was done under research permits no. 2013-251-ER-2014-177 (FM), 2013-256-NA-
338 2014-177 (JvR) and 2013-251-NA-2014-177 (OS) from the Tanzania Commission for Science and
339 Technology (COSTECH).

340 **Data availability:** Mitochondrial sequences are available on GenBank (accession numbers MH717378
341 - MH717444). Phenotypic data are available on Dryad (<https://doi.org/10.5061/dryad.v24tb57>).

342 **Competing interests.** We have no competing interests.

343 **Authors' contributions.** FM carried out fieldwork, participated in the design of the study, helped
344 generating stable isotope, morphological and genetic data, analysed the data and wrote the
345 manuscript together with OS. JvR and BN carried out fieldwork. SM helped in the lab and generated
346 the genetic data. OS designed and coordinated the study, made the sampling design, identified the
347 samples and wrote the manuscript together with FM.

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352 morphological measurements.

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354 to OS.

355

356 **Table 1:** Differentiation between the different species within Lake Chala in morphology (lower left) and
 357 trophic ecology (upper right), measured as the Bhattacharyya-distance in the morphospace spanned
 358 by PC1 and PC2 and the stable isotope space respectively. Significant differentiation is indicated by
 359 asterisks (*: $p < 0.05$, **: $p < 0.01$).

	<i>Astatotilapia</i>	<i>O. sp. "blue head"</i>	<i>O. hunteri</i>	<i>C. rendalli</i>
<i>Astatotilapia</i>		3.79**	1.98**	1.59**
<i>O. sp. "blue head"</i>	4.98**		0.28**	0.36**
<i>O. hunteri</i>	12.90**	0.90**		0.44**
<i>C. rendalli</i>	1.56**	4.97**	21.34**	

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364 **Table 2:** Coefficients for the morphological traits in the pairwise LDA. Coefficients higher than 10 or
 365 lower than -10 are highlighted bold (HUN = *O. hunteri*, BLU = *O. sp. "blue head"*, JIP = *O. jipe*, PAN = *O.*
 366 *pangani*). The last row indicates which species has the more positive LD1.

	HUN vs BLU	HUN vs JIP	HUN vs PAN	PAN vs JIP
Head length	22.6	38.1	40.2	24.9
Head width	5.4	17.8	7.9	5.1
Body depth	-10.5	-9.8	29	-7.6
Lower jaw length	3	2.4	9.6	1.6
Lower jaw width	1.7	-1.0	-2.1	5.5
Snout length	-3.5	-9.4	-23.3	-10.3
Snout width	-19.6	-2.9	-44.1	47.6
Cheek depth	-3.5	1.6	0.2	3.9
Preorbital depth	5.4	4.7	11.3	0.9
Interorbital width	-0.6	-9.6	-11.7	-35.3
Eye length	1.9	-11.8	5.5	-19.2
Eye depth	-0.1	-14.8	-28	6.4
Preorbital width	-14.1	-11.9	-0.9	-19.2
Positive LD1	HUN	HUN	HUN	PAN

367
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371 **Table 3:** Morphological differentiation between the different sampled *Oreochromis* species in the
 372 Nyumba ya Mungu Reservoir and its inflow, the Ruvu River, measured as the Bhattacharyya-distance
 373 in the morphospace spanned by PC1 and PC2. Significant differentiation is indicated by asterisks (*:
 374 $p < 0.05$, **: $p < 0.01$).

	<i>Astatotilapia</i>	<i>O. sp. "blue head"</i>	<i>O. jipe</i>	<i>O. pangani</i>	<i>O. esculentus</i>
<i>Astatotilapia</i>					
<i>O. sp. "blue head"</i>	14.37**				
<i>O. jipe</i>	18.96**	2.63**			
<i>O. pangani</i>	16.44**	3.13**	0.59		
<i>O. esculentus</i>	53.94**	16.26**	2.93**	11.11**	
<i>O. niloticus</i>	18.98**	2.11**	1.20**	5.46**	3.54**

375
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377

378 **Table 4:** Differentiation between the different endemic species of the Pangani/Ruvu drainage including
 379 Lake Chala in morphology, measured as the Bhattacharyya-distance in the morphospace spanned by
 380 PC1 and PC2. Significant differentiation is indicated by asterisks (*: p<0.05, **: p<0.01).

	<i>O. hunteri</i>	<i>O. jipe</i>
<i>O. hunteri</i>		
<i>O. jipe</i>	0.76**	
<i>O. pangani</i>	0.98*	0.59

381
 382
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 384

385 **Table 5:** Morphospace occupation after correcting for differences in sample sizes for the three
 386 focal species of this study and p-values for the difference in morphospace occupation between
 387 species.

Morphospace occupation			
	<i>O. hunteri</i>	<i>O. sp. "blue head" (Chala)</i>	<i>O. jipe</i>
<i>O. hunteri</i>	2.19		
<i>O. sp. "blue head" (Chala)</i>	3.23	> 0.20	
<i>O. jipe</i>	2.32	> 0.20	> 0.20
<i>O. sp. "blue head" (Ruvu)</i>	0.80	< 0.01	< 0.01

388

389 **Fig. 1** Map of the upper Pangani River drainage with the sampling sites used in this study: Lake Chala, the Pangani
390 Reservoir Nyumba ya Mungu and its main inlet, the Ruvu River, and Lake Babati

391

392 .

393 **Fig. 2** The oreochromine cichlid species of the upper Pangani system. From left to right and top to bottom:
394 *Oreochromis hunteri* from Lake Chala, adult male and subadult male; *O. jipe* from Nyumba ya Mungo, two adult
395 males; *O. pangani* from the Ruvu River, two subadults; *O. sp.* “blue head” from Lake Chala, adult male with
396 enlarged jaws; adult individual without enlarged jaws; *O. niloticus* from the Ruvu River (subadult); *O. sp.* adult
397 male from Lake Babati.

398

399

400 **Fig. 3** The phylogeographic relationships of the *Oreochromis* spp of Lake Chala, the upper Pangani and Lake
401 Babati. TCS haplotype network of 347 bp of mitochondrial D-loop. We only used 347 bp because most sequences
402 used from GenBank were restricted to this shorter fragment. The size of the circles indicates the number of
403 individuals with a given haplotype in our data, ticks between the circles indicate single point mutations. Dark
404 blue are individuals collected from Lake Chala, black individuals collected from the Nyumba ya Mungu Reservoir,
405 green individuals collected from the Ruvu River (upper Pangani, inflow to the Nyumba ya Mungu Reservoir),
406 turquoise individuals collected in Lake Babati, brown individuals were taken from GenBank (Nagl et al 2001,
407 D'Amato et al 2007, Nyingi et al 2009, Kinaro et al 2016). Whereas *O. hunteri* is very closely related to the upper
408 Pangani endemics *O. pangani* and *O. jipe*, we observed two only distantly related haplotype groups within *O. sp*
409 “blue head”, the other *Oreochromis* from Lake Chala and the Ruvu River. One of these is shared with the Rufiji-
410 Tilapia (*O. urolepis urolepis*).

411

412

413 **Fig. 4** Relative abundance of the different cichlid species at different water depths in Lake Chala. All four cichlid
414 species were present at all three geographical sampling locations. All species occurred from the very shallow
415 areas down to a water depth of at least 40 m but none was found deeper than 47 m. *Astatotilapia* were the most
416 abundant taxon in the shallow habitat, whereas *Coptodon rendalli* was most abundant in the deeper habitats.
417 The two *Oreochromis* species were both caught in the deep and in the shallow habitat, but were more common

418 in the deeper habitat and virtually absent from intermediate water depths. The endemic *Oreochromis hunteri*
419 was clearly the least abundant species in our catches. Abundance data is probably slightly biased towards the
420 smaller *Astatotilapia* and subadult *Oreochromis* and *Coptodon*, because we used a maximal mesh size of 22 mm
421 and therefore likely missed very large individuals of *Oreochromis* and *Coptodon*. Sample sizes are given to the
422 right of each depth category.

423

424

425 **Fig. 5** Ecological variation and differentiation among the four cichlid species in Lake Chala represented by the
426 stable isotope plot of the four species: the endemic *Oreochromis hunteri* (green), *O. sp.* “blue head” (blue),
427 *Coptodon rendalli* (black) and *Astatotilapia cf. bloyeti* (red). The size of the symbols corresponds to the size of
428 each specimen.

429

430

431 **Fig. 6** Morphological variation and differentiation among the cichlid species in Lake Chala and the Pangani
432 drainage. a) PC-plot on linear morphological traits with all individuals from Lake Chala (triangles) highlighted. (b)
433 Same morphospace with all fish from the Ruvu River (diamonds) and the Nyumba ya Mungu Reservoir (dots)
434 highlighted (b). (c) Morphospace occupation of the endemic species of the Pangani drainage and Lake Chala *O.*
435 *jipe* (yellow) from the Nyumba ya Mungu, *O. pangani* (orange) from the Ruvu River and *O. hunteri* from Lake
436 Chala (green) and (d) morphospace occupation of the introduced *Oreochromis sp.* “blue head” in Lake Chala
437 (blue) and in the Ruvu River (lightblue). The colour of the dots represent the different species with *Astatotilapia*
438 *cf. bloyeti* in red, *O. hunteri* in green, *O. sp.* ‘blue ‘head’ in blue, *C. rendalli* in black, *O. pangani* in orange, *O. jipe*
439 in yellow, *O. niloticus* in purple and *O. esculentus* in brown. The symbols of the dots represents the catching
440 location with triangle for Lake Chala, dots for the Nyumba ya Mungu Reservoir and diamonds for the Ruvu River.

441

442

443 **Fig. 7** Underwater picture of Lake Chala. (a) Picture from 2014 of all three invasive species at 6m water depth.
444 Note the high density of *Astatotilapia cf. bloyeti* (red) and *Coptodon rendalli* (black) in this habitat. (b) A school
445 of *O. hunteri* swimming in the open water in 1995.

446 References

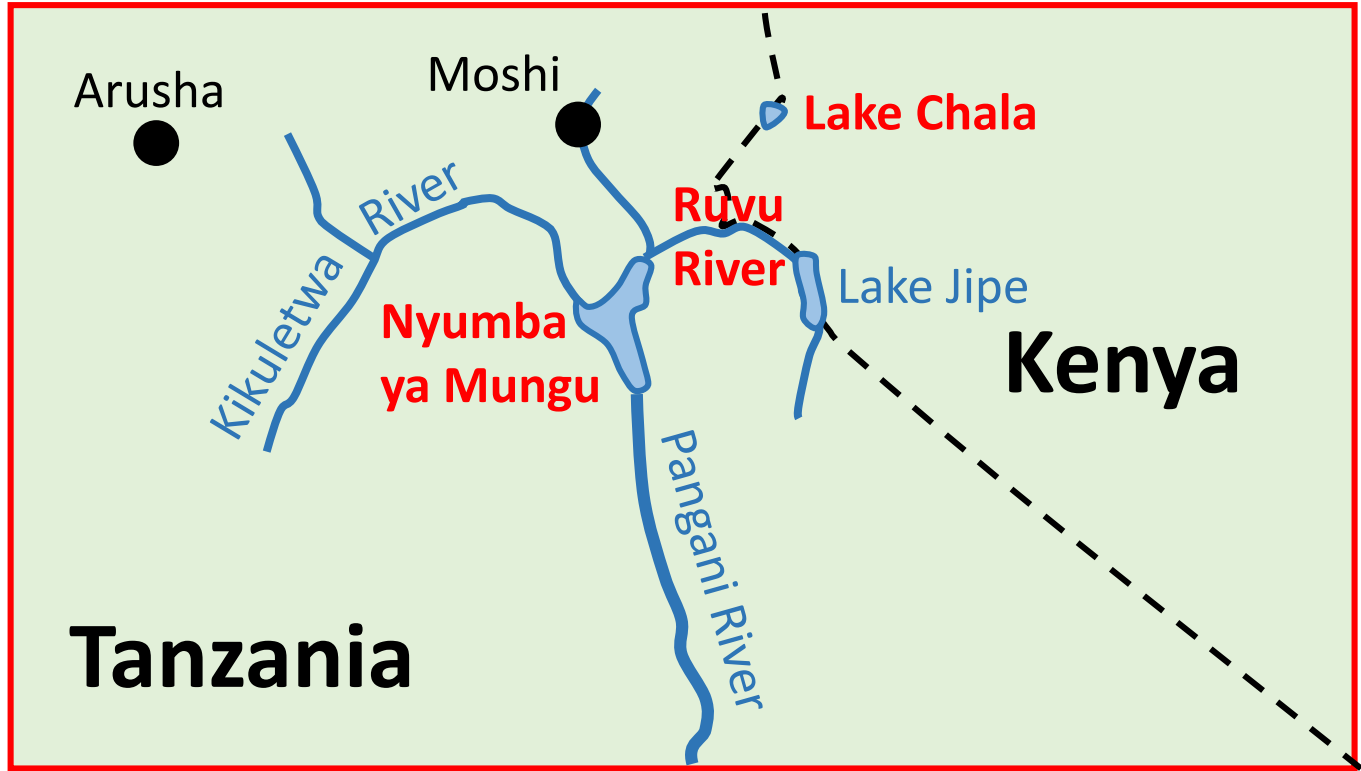
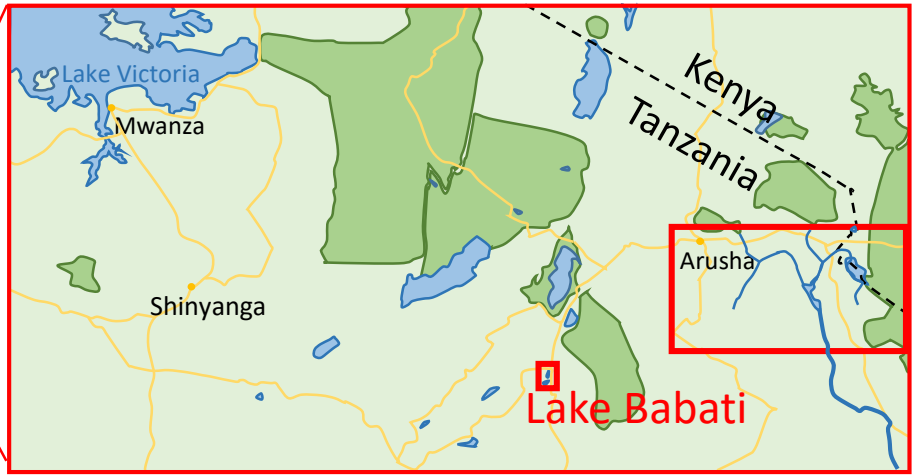
- 447 Angienda, P. O., H. J. Lee, K. R. Elmer, R. Abila, E. N. Waindi & A. Meyer, 2011. Genetic structure and
448 gene flow in an endangered native tilapia fish (*Oreochromis esculentus*) compared to invasive Nile
449 tilapia (*Oreochromis niloticus*) in Yala swamp, East Africa. *Conservation Genetics* 12: 243-55.
- 450 Bandelt, H. J., P. Forster & A. Rohlf, 1999. Median-joining networks for inferring intraspecific
451 phylogenies. *Molecular Biology and Evolution* 16: 37-48.
- 452 Casal, Christine Marie V., 2006. Global Documentation of Fish Introductions: the Growing Crisis and
453 Recommendations for Action. *Biological Invasions* 8: 3-11.
- 454 Crutchfield, J. U., 1995. Establishment and expansion of redbelly Tilapia and blue Tilapia in a power
455 plant cooling reservoir. *Uses and Effects of Cultured Fishes in Aquatic Ecosystems* 15: 452-61.
- 456 D'Amato, M. E., M. M. Esterhuyse, B. C. W. van der Waal, D. Brink & F. A. M. Volckaert, 2007.
457 Hybridization and phylogeography of the Mozambique tilapia *Oreochromis mossambicus* in southern
458 Africa evidenced by mitochondrial and microsatellite DNA genotyping. *Conservation Genetics* 8: 475-
459 88.
- 460 Dadzie, S., R. D. Haller & E. Trewavas. 1988. A Note on the Fishes of Lake Jipe and Lake Chale on the
461 Kenya-Tanzania Border
- 462 Daga, V. S., T. Debona, V. Abilhoa, E. A. Gubiani & J. R. S. Vitule, 2016. Non-native fish invasions of a
463 Neotropical ecoregion with high endemism: a review of the Iguacu River. *Aquatic Invasions* 11: 209-
464 23.
- 465 de Azevedo-Santos, V. M., O. Rigolin-Sa & F. M. Pelicice, 2011. Growing, losing or introducing? Cage
466 aquaculture as a vector for the introduction of non-native fish in Furnas Reservoir, Minas Gerais,
467 Brazil. *Neotropical Ichthyology* 9: 915-19.
- 468 DeGraaf, G. & L. Garibaldi, 2014. The value of african fisheries. *FAO Fisheries and Aquaculture*
469 *Circular* 1093
- 470 Deines, A. M., I. Bbole, C. Katongo, J. L. Feder & D. M. Lodge, 2014. Hybridisation between native
471 *Oreochromis* species and introduced Nile tilapia *O. niloticus* in the Kafue River, Zambia. *African*
472 *Journal of Aquatic Science* 39: 23-34.
- 473 Deines, A. M., M. E. Wittmann, J. M. Deines & D. M. Lodge, 2016. Tradeoffs among Ecosystem
474 Services Associated with Global Tilapia Introductions. *Reviews in Fisheries Science & Aquaculture* 24:
475 178-91.
- 476 Dieleman, J., M. Muschick, W. D. Nyingi & D. Verschuren, 2018. Species integrity and origin of
477 *Oreochromis hunteri* (Pisces: Cichlidae), endemic to crater Lake Challa (Kenya-Tanzania).
478 *Hydrobiologia*
- 479 Dieleman, J., B. Van Bocxlaer, C. Manntschke, D. W. Nyingi, D. Adriaens & D. Verschuren, 2015.
480 Tracing functional adaptation in African cichlid fishes through morphometric analysis of fossil teeth:
481 exploring the methods. *Hydrobiologia* 755: 73-88.
- 482 FAO, 2007. The state of world fisheries and aquaculture 2006. ed. FAO/OtU Nations. Rome, Italy
- 483 Firmat, C., P. Alibert, M. Losseau, J. F. Baroiller & U. K. Schliewen, 2013. Successive Invasion-
484 Mediated Interspecific Hybridizations and Population Structure in the Endangered Cichlid
485 *Oreochromis mossambicus*. *Plos One* 8
- 486 Ford, A. G. P., K. K. Dasmahapatra, L. Ruber, K. Gharbi, T. Cezard & J. J. Day, 2015. High levels of
487 interspecific gene flow in an endemic cichlid fish adaptive radiation from an extreme lake
488 environment. *Molecular Ecology* 24: 3421-40.
- 489 Genner, M. J., E. Connell, A. Shechonge, A. Smith, J. Swanstrom, S. Mzighani, A. Mwijage, B. P.
490 Ngatunga & G. F. Turner, 2013. Nile tilapia invades the Lake Malawi catchment. *African Journal of*
491 *Aquatic Science* 38: 85-90.
- 492 Genner, M. J., O. Seehausen, D. F. R. Cleary, M. E. Knight, E. Michel & G. F. Turner, 2004. How does
493 the taxonomic status of allopatric populations influence species richness within African cichlid fish
494 assemblages? *Journal of Biogeography* 31: 93-102.
- 495 Kinaro, Z. O., L. Y. Xue & J. A. Volatiana, 2016. Complete mitochondrial DNA sequences of the Victoria
496 tilapia (*Oreochromis variabilis*) and Redbelly Tilapia (*Tilapia zilli*): genome characterization and
497 phylogeny analysis. *Mitochondrial DNA* 27: 2455-57.

498 Lowe, M. R., W. Wu, M. S. Peterson, N. J. Brown-Peterson, W. T. Slack & P. J. Schofield, 2012.
499 Survival, Growth and Reproduction of Non-Native Nile Tilapia II: Fundamental Niche Projections and
500 Invasion Potential in the Northern Gulf of Mexico. *Plos One* 7
501 Lowe, R. H. , 1955. New species of Tilapia (Pisces, Cichlidae) from Lake Jipe and the Pangani River in
502 East Africa. *Bulletin of British Museum (Natural History)* 2: 19.
503 Moser, F. N., J. C. van Rijssel, S. Mwaiko, J. I. Meier, B. Ngatunga & O. Seehausen, 2018. The onset of
504 ecological diversification 50 years after colonization of a crater lake by haplochromine cichlid fishes.
505 *Proceedings of the Royal Society B-Biological Sciences* 285
506 Nagl, S., H. Tichy, W. E. Mayer, I. E. Samonte, B. J. McAndrew & J. Klein, 2001. Classification and
507 phylogenetic relationships of African tilapiine fishes inferred from mitochondrial DNA sequences.
508 *Molecular Phylogenetics and Evolution* 20: 361-74.
509 Nyingi, D., L. De Vos, R. Aman & J. F. Agnese, 2009. Genetic characterization of an unknown and
510 endangered native population of the Nile tilapia *Oreochromis niloticus* (Linnaeus, 1758) (Cichlidae;
511 Teleostei) in the Lobo Swamp (Kenya). *Aquaculture* 297: 57-63.
512 Ogutu-Ohwayo, R., R. E. Hecky, A.S. Cohen & L. Kaufman, 1997. Human impacts on the African Great
513 Lakes. *Environmental Biology of Fishes* 50: 117-31.
514 R-Core-Team, 2015. R: A Language and Environment for Statistical Computing.
515 Schliewen, U., D. Tautz & S. Pääbo, 1994. Sympatric speciation suggested by monophyly of crater lake
516 cichlids. *Nature* 368: 306-08.
517 Seehausen, O., 2007. Evolution and ecological theory - Chance, historical contingency and ecological
518 determinism jointly determine the rate of adaptive radiation. *Heredity* 99: 361-63.
519 Shuai, Fangmin, Xinhui Li, Yuefei Li, Li Jie, Yang Jiping & Sovan Lek, 2015. Forecasting the invasive
520 potential of Nile tilapia (*Oreochromis niloticus*) in a large subtropical river using a univariate
521 approach. *Fundamental and Applied Limnology / Archiv für Hydrobiologie* 187: 165-76.
522 Singh, A. K., D. Kumar, S. C. Srivastava, A. Ansari, J. K. Jena & U. K. Sarkar, 2013. Invasion and impacts
523 of alien fish species in the Ganga River, India. *Aquatic Ecosystem Health & Management* 16: 408-14.
524 Trewavas, E. 1983. *Tilapiine fishes of the genera Sarotherodon, Oreochromis and Danakilia*. London:
525 British Musuem (Natural Hististory).
526 Trewavas, E., J. Green & S. A. Corbet, 1972. Ecological Studies on Crater Lakes in West Cameroon
527 Fishes of Barombi Mbo. *Journal of Zoology* 167: 41-&.
528 Wagner, C. E., L. J. Harmon & O. Seehausen, 2014. Cichlid species-area relationships are shaped by
529 adaptive radiations that scale with area. *Ecology Letters* 17: 583-92.
530 Watson, R.T., A.H. Zakri & et al., 2005. *Ecosystems and Human Well-Being: Synthesis*. Washington:
531 Island Press.
532 Zhang, X. F., X. Y. Mei & R. D. Gulati, 2017. Effects of omnivorous tilapia on water turbidity and
533 primary production dynamics in shallow lakes: implications for ecosystem management. *Reviews in*
534 *Fish Biology and Fisheries* 27: 245-54.

535

Figure 1

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O. hunteri adult



O. hunteri subadult



O. jipe adult



O. jipe adult



O. pangani subadult



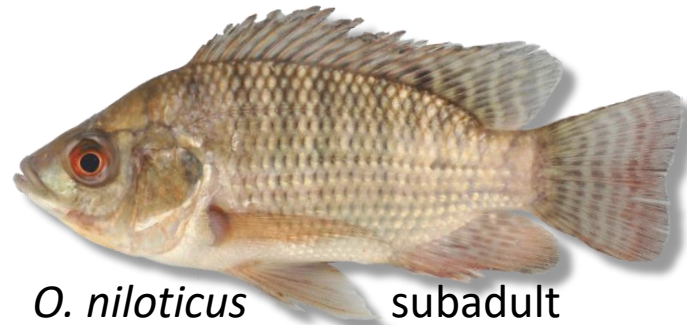
O. pangani subadult



O. sp. «blue head» adult



O. sp. «blue head» adult

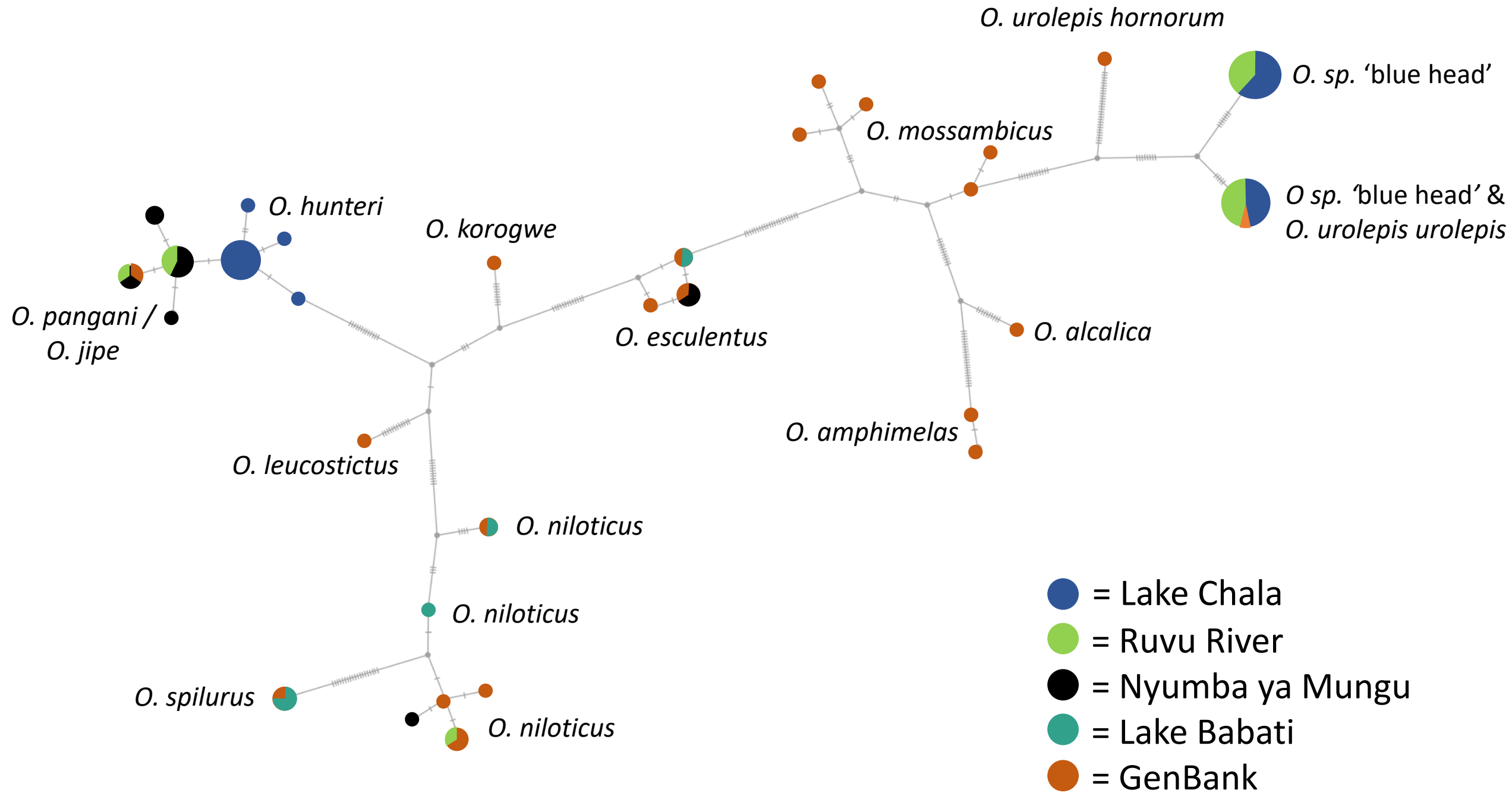


O. niloticus subadult



O. sp. «Babati» adult

Figure 3



Species abundance in Lake Chala

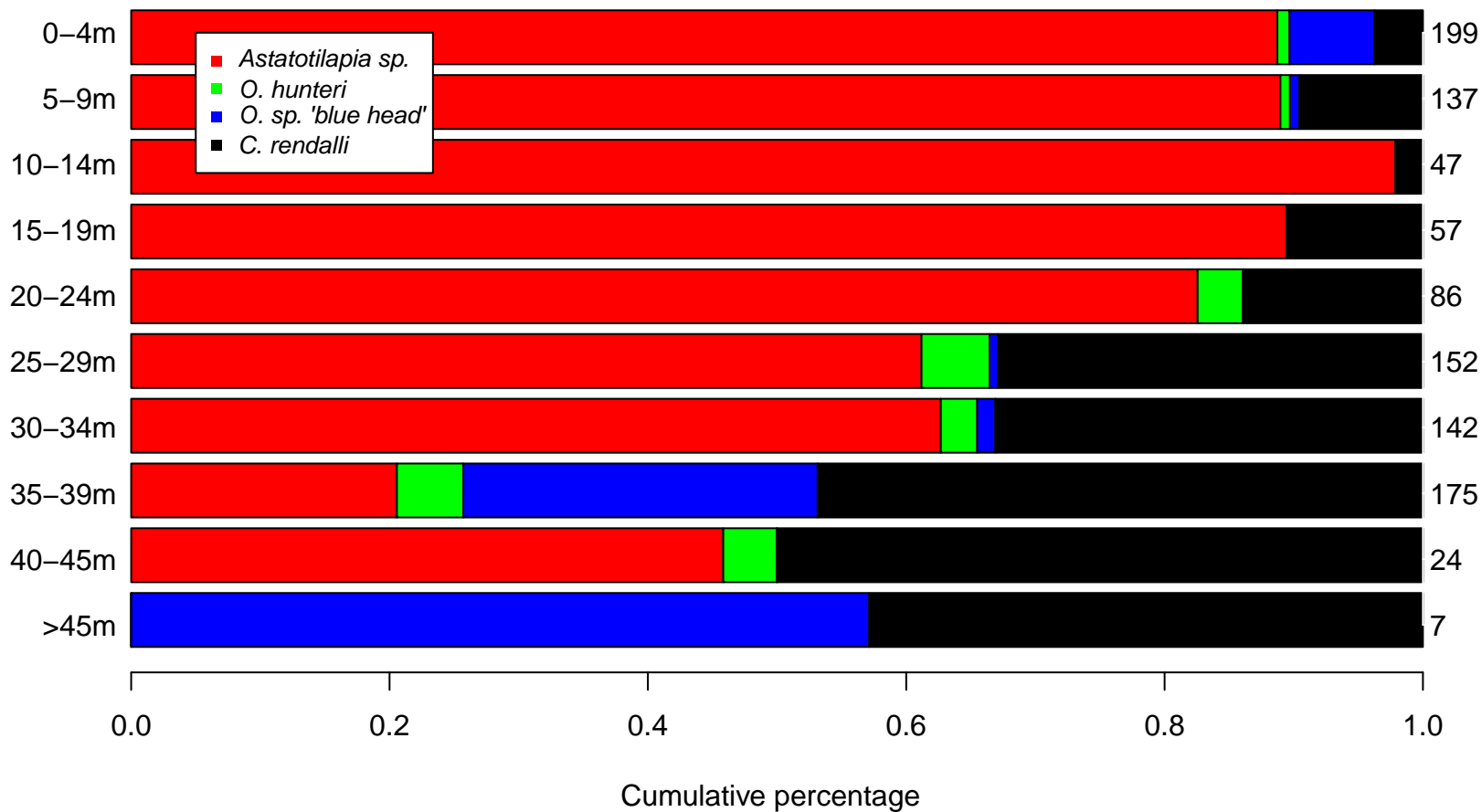


Figure 5

