1	Causes and mechanisms of synchronous succession trajectories in
2	primeval Central European mixed Fagus sylvatica forests
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15 Running headline

16 Synchronous succession in *Fagus sylvatica* forests

17 Abstract

1. Natural succession trajectories of Central European forest ecosystems are poorly 18 19 understood due to the absence of long-term observations and the pervasive effects of past 20 human impacts on today's vegetation communities. This knowledge gap is significant given that currently forest ecosystems are expanding in Europe as a consequence of global change. 21 2. Annually laminated sediments were extracted from two small lowland lakes (Moossee 521 22 m a.s.l.; Burgäschisee 465 m a.s.l.) on the Swiss Plateau. We combine high-resolution 23 24 palaeoecological and quantitative analyses to assess changes in vegetation during the Neolithic. We test for regionally synchronous land-use phases and plant successional patterns 25 that may originate from complex interactions between human and climatic impacts. 26

27 3. Mixed Fagus sylvatica forests dominated the Swiss Plateau vegetation over millennia. During the period 6,500–4,200 cal year BP, pronounced forest disruptions accompanied by 28 29 increased fire and agricultural activities occurred at c. 6,400–6,000 cal year BP, 5,750–5,550 cal year BP, around 5,400 cal year BP and at 5,100-4,600 cal year BP. Biodiversity increased 30 during these land-use phases, likely in response to the creation of new open habitats. After 31 32 decades to centuries of land-use, arboreal vegetation re-expanded. In a first succession stage, heliophilous Corylus avellana shrubs were replaced by pioneer Betula trees. These open 33 arboreal communities were outcompeted within 150-200 years by late successional F. 34 35 sylvatica and Abies alba forests. Most strikingly, cross-correlations show that these successions occurred synchronously (±11 years) and repeatedly over large areas (>1,000 km²) 36 and millennia. 37

4. *Synthesis.* First notable human impact shaped the primeval mixed *F. sylvatica* forests in Central Europe from *c.* 6,800–6,500 cal year BP on. Agrarian societies were susceptible to climate changes and we hypothesize that climate-induced, simultaneous agricultural expansion and contraction phases resulted in synchronous regional forest successions.
Currently forests are expanding in Central Europe as a result of land abandonment in marginal
areas. Our results imply that mixed *F. sylvatica* forests with *A. alba* and *Quercus* may re-expand
rapidly in these areas, if climate conditions will remain within the range of the mid-Holocene
climatic variability (with summers *c.* +1–2° C warmer than today).

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47 KEYWORDS

Biodiversity, Charcoal, Climate change, Cross-correlations, Fire ecology, Forest resilience,
 Non-pollen palynomorphs, Palaeoecology and land-use history, Pollen, Swiss Plateau

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51 **1. INTRODUCTION**

European forest ecosystems have expanded over the last decades in response to global 52 53 change involving increasing land abandonment in marginal regions (FOREST EUROPE, 2015). Most forests are still intensively managed and natural succession is restricted to a few 54 remnant areas (Ellenberg, 2009; Kaufmann, Hauck, & Leuschner, 2018). The direction of 55 ongoing forest change is difficult to assess since the projections are mostly based on short-56 term observations alone (Willis & Birks, 2006). After disturbance, forest vegetation 57 spontaneously tends towards a late successional stage, which is often unknown because 58 59 modern forests have been strongly altered by humans over centuries to millennia (Ellenberg, 2009; Puhe & Ulrich, 2001). Studying long-term successional patterns in the past (>200 years) 60 provides unique information about vegetation dynamics under quasi-natural conditions, that 61 62 may help assess future vegetation dynamics (Foster, Schoonmaker, & Pickett, 1990). Specifically, the assessment of vegetation shifts at the onset of the Neolithic, when humans 63 gradually became sedentary in Europe (Guilaine, 2015; Müller, 2015; Schier, 2015). This period 64

provides insights into successional patterns of moderately disturbed forests that were only
marginally shaped by human activities under environmental conditions comparable to today
(Kalis, Merkt, & Wunderlich, 2003). That information is crucial for nature conservation and
forest management in protected areas such as national parks (Feurdean & Willis, 2008;
Lindbladh, Niklasson, Karlsson, Björkman, & Churski, 2008; Valsecchi, Carraro, Conedera, &
Tinner, 2010).

71 Previous palaeoecological studies in Central Europe have nicely illustrated that early 72 successional (e.g. pioneer herbs, Corylus avellana, Betula) and late successional stages (Fagus sylvatica, Abies alba) oscillated more frequently and pronouncedly after 6,500 cal year BP 73 (Ammann, 1989; Becker et al., 2006; Kleinmann, Merkt, & Müller, 2015; Rey et al., 2017). 74 75 Statistical analyses confirmed that prominent changes in vegetation were directly linked to 76 land-use (Gobet, Tinner, Hochuli, van Leeuwen, & Ammann, 2003; Rey et al., 2013; Schwörer, Colombaroli, Kaltenrieder, Rey, & Tinner, 2015; Tinner, Hubschmid, Wehrli, Ammann, & 77 Conedera, 1999). Fires were commonly used to open landscapes to create space for 78 79 settlements, pastoral and arable land (Jacomet et al., 2016; Tinner, Conedera, Ammann, & Lotter, 2005), primarily disadvantaging late-successional trees and favouring the expansion of 80 81 heliophilous herbs and shrubs (Conedera, Colombaroli, Tinner, Krebs, & Whitlock, 2017; 82 Tinner et al., 2000). Many studies numerically show that land-use substantially increased open 83 land biodiversity over the long-term (Colombaroli, Beckmann, van der Knaap, Curdy, & Tinner, 2013; Colombaroli & Tinner, 2013; Giesecke, Ammann, & Brande, 2014). However, one crucial 84 question still remains: were the marked vegetation changes spatially synchronous? This 85 86 question is important to assess the underlying causes of ecosystem change. Synchronism 87 would speak for a superimposed forcing such as climate (i.e. precipitation and/or temperature 88 changes) or, perhaps less likely, concerted cultural activities (Berglund, 2003; Tinner et al.,

89 2003), whereas diachronism would rather point to spatio-temporally randomized land-use phases of local autarchic societies that were acting independently from each other (Rösch & 90 Lechterbeck, 2016). Until now, it was impossible to test these competing hypotheses because 91 92 available records lacked temporal precision and resolution (Berglund, 2003; Tinner et al., 93 2003). In this study, we aim to resolve this long-standing question with high-resolution 94 palaeoecological and time-series analyses that rely on varved sediment chronologies reaching 95 exceptional temporal precision of c. ±20–50 cal years (Rey et al., 2018). The overarching goal of the study is to explore the nature of re-occurring broad-scale successional patterns in mixed 96 97 beech forests after moderate human impact to refine existing projections of future forest 98 transformations (Bugmann et al., 2015; Ruosch et al., 2016) under global change conditions that are currently triggering forest expansions in Central Europe and adjacent regions. 99

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101 2. MATERIAL AND METHODS

102 **2.1. Study sites**

103 The study area is located on the western Swiss Plateau in southern Central Europe (Figure 1). Moossee is a small lake close to the city of Bern at 521 m a.s.l. (47°1'17.0"N, 7°29'1.7"E). The 104 105 lake has a maximum water depth of 22 meters and a surface area of 0.31 km². Burgäschisee 106 is a small lake further to the northeast (c. 25 km from Moossee) at 465 m a.s.l. (47°10'8.5"N, 7°40′5.9″E). It has a maximum water depth of 31 meters and a surface area of 0.21 km². Both 107 108 lakes are eutrophic today with anoxic waters in the hypolimnion (Guthruf, Zeh, & Guthruf-109 Seiler, 1999). The present climate is temperate with mean annual temperatures of 8.8°C and an annual rainfall of 1,059 mm at Moossee (data from Bern/Zollikofen at c. 3 km distance, 110 111 MeteoSwiss, 2017). At Burgäschisee, it is slightly warmer and wetter with mean annual 112 temperatures of 9.1°C and an annual rainfall of 1,088 mm (data from Koppigen, at c. 6 km distance, MeteoSwiss, 2017). The current vegetation at both lakes is dominated by mixed *F*. *sylvatica* forests on the surrounding hills and scattered mixed *Alnus glutinosa-Fraxinus excelsior* stands on the lake shores. Most of the flat areas around the lakes are intensively
used for agriculture. The region has been important for agricultural activities for millennia as
indicated by frequent local archaeological findings dating back to at least 6,800-6,500 cal year
BP (Hafner, Harb, Amstutz, Francuz, & Moll-Dau, 2012; Harb, 2017; Wey, 2012).

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120 **2.2. Corings and chronologies**

In 2014, three parallel sediment cores at Moossee and two parallel sediment cores at
Burgäschisee were retrieved with a UWITEC piston corer (core diameter: 9 cm, core length:
200 cm) in deep parts of the lakes. A total of 7 m of lake sediments were sampled at each site
covering *c*. 9,000 years. For this study, we focus on the Neolithic to Early Bronze Age period
(6,500–3,800 cal year BP) during which the sediments are annually laminated (i.e. varved) at
both lakes (see Rey et al., 2018).

127 The Moossee chronology is based on 27 terrestrial plant macrofossils and varve counts (Rey et al., 2018). The program OxCal 4.3 (V-sequence; Bronk Ramsey, 1994, 1995, 2001; Bronk 128 129 Ramsey, van der Plicht, & Weninger, 2001) and the IntCal13 calibration curve (Reimer et al., 130 2013) were used to estimate the age-depth model (Figure 2a). The Burgäschisee chronology 131 is based on 22 terrestrial plant macrofossils and basic assumptions about the sedimentation rates from a previous chronology (Rey et al., 2017, 2018). Here, the program OxCal 4.3 (U/P-132 133 sequences; Bronk Ramsey, 1994, 1995, 2001, 2008; Bronk Ramsey & Lee, 2013) and the IntCal13 calibration curve (Reimer et al., 2013) were used for the estimations of the age-depth 134 135 model (Figure 2b). Both records indicate fairly regular sedimentation rates for the investigated 136 time interval with an average of 26 years/cm at Moossee and 18.7 years/cm at Burgäschisee.

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138 **2.3.** Pollen, macrofossil, non-pollen palynomorphs and charcoal analyses

All palaeoecological analyses were conducted on the same cores and for the same sample 139 140 depths (Moossee: 312 samples, Burgäschisee: 308 samples). We applied a contiguous, approximately constant year sampling $(10 \pm 2 \text{ years/sample})$. For all microscopic analyses 141 (pollen, non-pollen palynomorphs (NPPs), microscopic charcoal), samples of 1 cm³ were 142 treated with HCl, KOH, HF, acetolysis, sieved with a mesh size of 500 µm and mounted in 143 glycerine following standard palynological methods (Moore, Webb, & Collinson, 1991). 144 145 Lycopodium tablets (University of Lund batch no. 1031 with 20,848 ± 3,457 spores per tablet 146 for Moossee and University of Lund batch no. 177745 with 18,584 ± 1,853 spores per tablet for Burgäschisee) were added prior to the chemical treatment to estimate microfossil 147 concentrations (Stockmarr, 1971). We used palynological keys (Beug, 2004; Moore, Webb, & 148 149 Collinson, 1991), photo atlases (Reille, 1992) and the reference collection at the Institute of 150 Plant Sciences (University of Bern) to identify pollen and spores under a light microscope at 151 400× magnification. Phase contrast under 1,000× magnification was used to separate various Cerealia-types (Hordeum-type, Triticum-type; Beug, 2004). The minimum number of 152 terrestrial pollen grains counted per sample was 500 with a total of 132 (Moossee) and 113 153 154 (Burgäschisee) identified terrestrial pollen types. The results are presented as percentages of 155 the terrestrial pollen sum (Figures 3 and 4), in which we excluded spores, pollen of aquatic plants and other microfossils (e.g. stomata, algae cells). Pollen can be dispersed by wind over 156 157 large distances (up to 50 km) and generally represents extra-local to regional vegetation dynamics (Conedera, Tinner, Crameri, Torriani, & Herold, 2006). However, since the sites are 158 rather small (Moossee: 0.31 km², Burgäschisee 0.21 km²), we assume that the pollen source 159

area only spans several hundred meters to at most a few kilometers (Conedera, Tinner,
Crameri, Torriani, & Herold, 2006; Sugita, 1994).

Samples for macrofossil analyses (including macroscopic charcoal) had mean volumes of 17.5 cm³ (Moossee) and 11 cm³ (Burgäschisee) and were sieved with a mesh size of 200 µm. Terrestrial remains (e.g. *F. sylvatica* bud scales) were identified under a stereo microscope. Macrofossil concentration (remains cm⁻³) were calculated and plotted in Figures 3 and 4. Plant macrofossils mainly indicate local vegetation as a result of short dispersal distances (1–100 m; Birks, 2003).

Microscopic charcoal particles >10 μm and <500 μm were counted on pollen slides
following Tinner and Hu (2003) and Finsinger and Tinner (2005) and presented as microscopic
charcoal influx values (particles cm⁻² year⁻¹; Figures 3 and 4). These values were used as a proxy
for regional fire activity since microscopic charcoal particles are dispersed over distances of *c*.
20-50 km (Adolf et al. 2018; Tinner et al. 1998).

Macroscopic charcoal pieces >600 μ m were identified under a stereo microscope during macrofossil analysis. Their occurrence (pieces cm⁻³) may primarily reflect local fire activity within few hundred meters distance (Adolf et al., 2018; Figures 3 and 4). This procedure (large sampling volume, size threshold >600 μ m) allows reducing the influence of regional fires on macroscopic charcoal, without using de-trending and peak isolation approaches (Adolf et al., 2018).

Green algae (*Botryococcus, Tetraedron, Coelastrum*) and akinetes of cyanobacteria (*Anabaena, Aphanizomenon*) were counted on pollen slides to estimate changes in the trophic levels. Although, single NPP taxa might occur under natural conditions, contemporaneous or consecutive high numbers of green algae and cyanobacteria are considered as a proxy for eutrophication during or after local anthropogenic impacts (van Geel, Mur, RalskaJasiewiczowa, & Goslar, 1994; Hillbrand, van Geel, Hasenfratz, Hadorn, & Haas, 2014). All NPPs
are presented as percentages of the terrestrial pollen sum (Figures 3 and 4).

Local pollen assemblage zones (LPAZ; Figures 3 and 4) were delimited using optimal sum-of-squares partitioning (Birks & Gordon, 1985) with the program ZONE 1.2 (Juggins, 1991). Statistically significant zones were assessed with the program BSTICK (Line & Birks, unpublished software) following the broken-stick method (Bennett, 1996). The programs Tilia 2.0.41 and CorelDraw were used to plot the data for both high-resolution sequences (Figures 3 and 4).

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193 **2.4. Biodiversity estimations**

194 Rarefaction analysis was used to calculate the palynological richness (PRI), which is commonly used as a proxy for species richness at local to regional scale in many palaeoecological studies 195 196 (e.g. Birks & Line, 1992; Colombaroli & Tinner, 2013; Odgaard, 1999; Schwörer, Colombaroli, 197 Kaltenrieder, Rey, & Tinner, 2015). Rarefaction analysis allows estimating the number of taxa 198 per sample for a constant counting sum (Birks & Line, 1992), which was 500 at both sites. The 199 probability of interspecific encounter (PIE; Hurlbert, 1971) was used as a measure of palynological evenness. In previous studies PIE was used to evaluate effects of few dominant 200 201 pollen producers (e.g. C. avellana; van der Knaap, 2009) on the palynological assemblages and 202 in particular on palynological richness. To assess such effects, PIE-derived evenness-detrended palynological richness (DE-PRI) was calculated following Colombaroli, Beckmann, van der 203 204 Knaap, Curdy, and Tinner (2013). This approach aims at removing the evenness trend from 205 palynological richness by building an ordinary least square regression (OLS) between the 206 dependent (PRI) and the independent variable (palynological evenness). The resulting 207 residuals of pollen richness (= PRI - PIE) are then distributed around the original pollen richness

values (Colombaroli & Tinner, 2013). Only if PRI and DE-PRI show similar changes and trends,
we assume that the species richness estimates are unaffected by evenness effects. The DEPRI approach has proven to be valuable to estimate contemporary species richness patterns
at the European scale (Adolf, 2017). The program R statistics (R Development Core Team,
2016) was used for all calculations (see results in Figures 3 and 4).

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214 **2.5.** Time-series analyses (cross-correlations)

215 Cross-correlations (Gobet, Tinner, Hochuli, van Leeuwen, & Ammann, 2003; Green, 1981; Rey 216 et al., 2013; Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999) were calculated to 217 identify leads and lags between fire (microscopic charcoal influx values) and vegetation (pollen percentages, PRI, PIE, DE-PRI) by using the program MYSTAT 12 (Systat, 2007). The time 218 window for all calculations was 6,500–4,150 cal year BP that included 209 samples at Moossee 219 220 (1 sample = 11.1 ± 0.2 cal years) and 266 samples at Burgäschisee (1 sample = 8.7 ± 0.6 cal 221 years). Cross-correlation coefficients for microscopic charcoal influx values vs. pollen 222 percentages, PRI, PIE and DE-PRI (see Figure 5) were calculated at ±50 lags corresponding to 223 ±555 cal years (Moossee) and ±435 cal years (Burgäschisee), following Bahrenberg, Giese, Mevenkamp, and Nipper (2008), with maximum lag numbers corresponding to one fourth of 224 225 the sample number N (lag number \leq N/4). The correlation coefficients at lag 0 between 226 microscopic charcoal and selected pollen types are represented as detailed correlograms (Figure 6). For cross-correlation analyses between the two sites (e.g. F. sylvatica percentages 227 228 at Moossee vs. F. sylvatica percentages at Burgäschisee), the time series with higher time 229 resolution (Burgäschisee) was adjusted to the one with lower resolution (Moossee) by 230 combining neighbouring samples following Tinner et al. (2015). This led to a total of 209 231 samples (1 sample = 1 lag = 11.1 ± 0.2 cal years) and ± 50 lags (= ± 555 cal years) that were

included in the cross-correlation calculations (Figure 7). Variables were not de-trended prior to the cross-correlation analyses because the vegetation type remains more or less uniform for the whole time window without pronounced population trends (see Figures 3 and 4; Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999). The 95% (2 σ) confidence intervals of the correlations were estimated by computing ±2 standard errors of the correlation coefficients (Bahrenberg, Giese, Mevenkamp, & Nipper, 2008). This corresponds to a twosided significance level (α) of 5% (Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999).

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240 **3. RESULTS**

3.1. Vegetational, agricultural and fire dynamics

242 Both studied pollen records (Moossee: 7,000–3,900 cal year BP, Burgäschisee: 6,500–3,800 243 cal year BP) have a high temporal resolution of c. 10 cal years per sample and an extraordinary chronological precision (Rey et al., 2018). Zonation analyses show a high number of significant 244 LPAZ (Moossee: 15, Burgäschisee: 14; see Figures 3 and 4), suggesting high variability of 245 246 vegetation composition over time. In general, F. sylvatica was dominant in the closed forests 247 (pollen percentages generally \geq 20%). This pollen-based interpretation is supported by regular finds of macrofossils (bud scales). Other important taxa in the forests included Quercus and A. 248 249 alba. The latter shows slightly higher pollen percentages at Burgäschisee (up to 10% instead 250 of 5%), probably as a result of increasing precipitation towards the east of the Swiss Plateau, 251 making the species more competitive under mesophilous conditions (Rey et al., 2017). However, macrofossil and stomata finds of A. alba point to local occurrences of the species at 252 both sites. Alnus glutinosa was preferably growing on the wet soils at the lake shores as 253 254 indicated by high and fairly stable pollen percentages (around 25%) and abundant 255 macrofossils finds (e.g. A. glutinosa fruits). Pollen percentages of other subordinate taxa such 256 as Tilia, Ulmus and Taxus baccata steadily decrease until c. 5,000 cal year BP after which they 257 remain low (<2%, see Ulmus) or even disappear (see Tilia and T. baccata), most likely as a 258 result of overexploitation and fire disturbance (Rey et al., 2017; Tinner, Conedera, Ammann, 259 & Lotter, 2005). Distinct phases with low F. sylvatica pollen percentages (<10%), increased 260 cultural indicators (Cerealia-type, *Plantago lanceolata, Linum usitatissimum*-type), apophytes 261 (Urtica), other herbs (Poaceae, Allium-type) and heliophilous shrubs (C. avellana, Juniperus) 262 are apparent at both sites. A first phase is dated at 6,550–6,150 cal year BP at Moossee and 263 at 6,300–5,950 cal year BP at Burgäschisee. Subsequently, similar phases occurred at 5,750– 5,550 cal year BP, at around 5,400 cal year BP and at 5,100–4,600 cal year BP at both sites. 264 Finally, a last phase is dated around 3,900 cal year BP at Moossee. Periods of forest disruption 265 266 were generally accompanied by high regional fire activity as suggested by increasing microscopic charcoal influx values (>5,000 particles cm⁻² yr⁻¹), sometimes overlapping with 267 local fire activity as inferred from macroscopic charcoal presence (≥600 µm). This finding is in 268 269 good agreement with local archaeological evidence (e.g. pile dwellings, light grey bars in 270 Figures 3 and 4). Some settlements around the lake had impacts on the water quality as shown 271 by high numbers of cyanobacteria (*Anabaena*, *Aphanizomenon*) and green algae (*Tetraedron*) 272 between 5,750 and 5,550 cal year BP at both lakes and around 4,600 cal year BP at Moossee. 273 Such blooms are best explained by strong lake water eutrophication. After the settlement 274 phases, Betula as a pioneer tree species quickly recolonized abandoned places (pollen 275 percentages up to 30% and abundant macrofossils). After intermediate phases often with F. 276 excelsior, Ulmus and Tilia expansions, these trees were unremittingly replaced within 150–200 277 years by late successional tree species (F. sylvatica, A. alba) if no further disturbances 278 occurred. These succession cycles were simultaneously repeated at both sites and are not only

visible in the pollen percentage data but are also confirmed by plant macrofossils, indicating
possible broad-scale succession trajectories.

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282 **3.2. Biodiversity reconstruction**

283 Palynological richness (PRI) and evenness-detrended palynological richness (DE-PRI) show 284 good agreements (Figures 3 and 4), suggesting that overall trends in palynological richness are not strongly affected by evenness. PRI and DE-PRI generally increase during phases with higher 285 286 human impact (>25 pollen types per sample), i.e. around 6,400 cal year BP, 4,600 cal year BP 287 and 3,900 cal year BP at Moossee, at 5,700–5,600 cal year BP at both sites, and around 4,900 288 cal year BP at Burgäschisee. These increases are directly related to vegetation openings and 289 the introduction of cultivated plants (*Hordeum*-type, *Triticum*-type, *L. usitastissimum*-type). 290 The establishment of open land for Neolithic farming created new habitats for short-lived 291 weeds (P. lanceolata), apophytes (Urtica) and light-demanding shrubs (C. avellana, Sambucus 292 nigra). Palynological evenness as inferred from PIE (Figures 3 and 4) is more or less stable (PIE 293 around 0.85). Slightly lower values of palynological evenness (PIE around 0.8) are recorded in 294 phases when either Betula or F. sylvatica pollen grains are dominant and the forests were 295 rather closed, suggesting that the expansion of monospecific stands may have affected species 296 evenness. On the other hand, phases with open forests (e.g. 5,700–5,600 cal year BP at both 297 sites) have usually high palynological evenness values (PIE ≥ 0.85), suggesting that vegetation evenness is correlated to openness. 298

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300 **3.3. Time-series analyses**

301 At both sites, pollen percentages of trees have significant negative correlations with 302 microscopic charcoal influx (maximum negative correlations at lag 0; Figure 5). *F. excelsior* is 303 a good example to illustrate immediate negative fire impacts. The negative correlations of F. 304 sylvatica pollen percentages is, however, markedly delayed at both sites (c. 320 years after a 305 fire at Moossee respectively c. 80 years after a fire at Burgäschisee), possibly indicating a slightly higher fire tolerance compared to other tree species (Tinner et al., 2000). On the other 306 307 hand, pollen percentages of herbs (e.g. Cerealia-type) reach maximum positive correlations 308 with microscopic charcoal at lag 0, showing that agropastoral activities were strongly 309 connected to the use of fire. Similarly, correlations between microscopic charcoal influx and 310 PRI, DE-PRI and at Moossee also PIE suggest that biodiversity increased in the new habitats (i.e. open areas) created by anthropogenic burning for agricultural purposes. Pollen 311 percentages of shrubs such as C. avellana and Salix also have significant positive correlations, 312 313 e.g. at Burgäschisee maximum positive correlations are reached at lag +4 for C. avellana (c. 35 314 years after a fire) and at lag 0 for Salix (Figure 5b). Betula shows maximum positive correlations at lag +35 (c. 390 years after a fire; Figure 5a) at Burgäschisee and at lag +41 (c. 355 years after 315 a fire; Figure 5b) at Moossee. This result with *C. avellana* preceding *Betula* pollen percentages 316 317 is likely caused by the successional trajectories after fire disturbance. Later on, the recovery 318 of F. excelsior follows (lag +41 and lag +45, corresponding to c. 455 and c. 390 years) after the 319 fire-related decline (negative correlations) as well as *F. sylvatica* (lag +50 = c. 435 years; Figure 320 5b). Correlation coefficients at lag 0 between microscopic charcoal influx values and pollen 321 types or PRI, PIE and DE-PRI show similar linkages (Figure 6), with significant positive 322 correlations for cultural indicators (Hordeum-type, Triticum-type), other herbs (e.g. Allium-323 type, Mentha-type, Papaver rhoeas-type, Urtica), PRI, DE-PRI and shrubs (e.g. C. avellana, 324 Salix, S. nigra). In contrast, most tree pollen types (e.g. Quercus, F. excelsior, F. sylvatica, 325 Ulmus, T. baccata) have significant negative correlations with microscopic charcoal, 326 suggesting that forest trees were most affected by anthropogenic burning. However, A.

glutinosa-type as an important tree pollen type is indifferent showing neither significant
 positive nor negative correlations with charcoal-inferred fire incidence.

If cross-correlations are used to compare the two sites, a striking pattern appears 329 (Figure 7). Total shrubs and trees as well as some taxa (e.g. C. avellana, Betula, F. sylvatica, 330 331 *Tilia*) show pronounced positive correlations with maximum correlations at lag 0, suggesting 332 synchronous oscillations within the pooled sample resolution (11.1 ± 0.2 cal years). Cross-333 correlations between pollen percentages of herbs (Moossee data vs. Burgäschisee data), F. excelsior and A. alba only slightly differ and may reflect local vegetation variability, however, 334 335 significant positive correlations also center around lag 0 with maximum positive correlations within c. ±35 years. Of particular interest are the correlations for the cereals, that also show 336 this pattern, confirming that the cultural phases at the two sites were largely coeval with only 337 338 minor differences, possibly indicating positive synergy effects of cereal production in old and new prehistoric centres. Similarly, also PRI and DE-PRI co-vary in time at the two sites. Taken 339 340 together, these results quantitatively show that the general succession patterns after disturbance were not only similar as presented in Figures 3 and 4 but in fact synchronous at 341 342 the two sites.

343

344 **4. DISCUSSION**

4.1. Impacts of land-use on temperate forests and synchronous patterns of forest succession after disturbances

The palaeobotanical data suggest the dominance of mixed beech forests over millennia at our two study sites Burgäschisee and Moossee. Land-use phases contributed to openings and a gradual conversion from mixed beech to almost pure beech forests. These general patterns have been observed elsewhere in the lowlands of southern Central Europe after *c*. 7,000 cal 351 year BP (e.g. Ammann, 1989; Kleinmann, Merkt, & Müller, 2015; Rösch & Lechterbeck, 2015; Tinner & Ammann, 2005; Tinner, Conedera, Ammann, & Lotter, 2005). However, local 352 differences in species composition resulted from dissimilar environmental conditions. For 353 instance in Switzerland, moisture availability increases towards higher elevations and to the 354 355 east. This moisture gradient may have led to the co-dominance of *A. alba* in the forests (Gobet 356 & Tinner, 2012; van der Knaap, van Leeuwen, & Ammann, 2004; Lotter, 1999; Wehrli, Tinner, 357 & Ammann, 2007), given that as the most shade tolerant and tallest tree species of Europe, 358 silver fir is particularly competitive under mesophilous conditions (Ellenberg, 2009; Ruosch et al., 2016; Tinner et al., 2013). Conversely, drier conditions in the Lake District (Seeland) in the 359 rain shadow of the Jura Mountains may have promoted Quercus spp. including 360 submediterranean Q. pubescens (Ammann, 1989; Hadorn, 1992; Steiger, 2010). 361

362 Several diebacks of mixed F. sylvatica forests occurred between 6,400 and 6,000 cal year BP, at 5,800–5,650 cal year BP, 5,400–5,300 cal year BP, at 5,100–4,700 cal year BP at 363 both sites and around 3,900 cal year BP at Moossee only. The combined charcoal and pollen 364 evidence as well as the cross-correlation analyses suggest that burning was used as a tool for 365 366 opening the mixed F. sylvatica forests, creating a mosaic of new habitats that advantaged 367 many herbaceous species, increasing the overall biodiversity (see biodiversity estimations in Figures 3 and 4). A similar pattern has been reconstructed in previous studies and might be 368 369 typical for European landscapes (Colombaroli, Beckmann, van der Knaap, Curdy, & Tinner, 2013; Colombaroli & Tinner, 2013; Giesecke, Ammann, & Brande, 2014). Prehistoric slash-and-370 371 burn activities were strong enough to release a wide-scale expansion of early-successional 372 shrublands, typically dominated by C. avellana, Betula and Alnus (Aaby, 1986; Tinner, 373 Conedera, Ammann, & Lotter, 2005; Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999). 374 In contrast, most tree species were strongly reduced by fires and other human disturbances

375 such as logging, pollarding and browsing (e.g. Haas, Karg, & Rasmussen, 1998; Haas & 376 Schweingruber, 1993; Pott, 1985; Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999). Most interestingly, some tree species such as F. sylvatica and A. alba seem to have fully 377 recovered and even expanded after forest disruptions (within c. 150-200 years after a land-378 379 use phase), suggesting long-term resilience to moderate human disturbance, while others 380 (Ulmus, Tilia, Acer and T. baccata) markedly diminished, suggesting high sensitivity of these 381 species. Ultimately, at c. 5,000 cal year BP (Figure 3 and 4) this process of repeated 382 disturbances led to a strong impoverishment in species diversity of forest ecosystems (Tinner & Ammann, 2005), generating the advent of monospecific forests that are characteristic for 383 Central European landscapes today (e.g. almost pure spruce, beech, larch, oak stands, 384 depending on altitude and other environmental conditions; Ellenberg, 2009; Gobet, Vescovi, 385 386 & Tinner, 2010).

Here, we quantitatively demonstrate for the first time that forest succession cycles 387 over 3,000 years were regionally synchronous (Figure 7). The distance between the two sites 388 (c. 25 km) exceeds the expected pollen catchment expected for the two study sites (most 389 390 pollen from within a few hundred meters to at most a few kilometers; Conedera, Tinner, 391 Crameri, Torriani, & Herold, 2006), arguing for coeval population dynamics over the millennia within the time resolution and precision available (c. 20-50 cal years; Rey et al., 2018). 392 393 Because this pattern was found in the cross-correlation analyses for several tree taxa as well as for the tree percentage sum, we consider it to be very reliable. Synchronous forests 394 395 disruptions (tree pollen <80%) may have occurred at several distant sites in Central and 396 Southern Europe (Figure 8). The few available high-resolution (c. 10-20 years) and high-397 precision (sufficient radiocarbon dates on terrestrial material) time series covering the entire 398 Neolithic and the onset of the Bronze Age reveal coeval tree population reductions at c. 6,400399 5,900 cal year BP, 5,750–5,600 cal year BP, 5,400 cal year BP, 5,100–4,700 cal year BP and 400 around 3,900 cal year BP (Ammann, 1989; Kleinmann, Merkt, & Müller, 2015; Lotter, 1999; 401 Rösch, Kleinmann, Lechterbeck, & Wick, 2014a, b; Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999; Wehrli, Tinner, & Ammann, 2007). One of these sites, Lago di Origlio, is 402 403 located in the lowlands south of the Alps (Sottoceneri), where environmental conditions are 404 typical for submediterranean Southern Europe (warmer summers, milder winters). We thus 405 hypothesize the existence of supra-regional synchronicity of forest disturbances during the 406 Neolithic, as it was formerly advocated for Central and Southern Europe (including Sicily) for the Bronze Age, the Iron Age, the Roman Period and the subsequent early medieval times 407 (Finsinger & Tinner, 2006; Gobet, Tinner, Hochuli, van Leeuwen, & Ammann, 2003; Tinner et 408 409 al., 2003, 2009).

410

411 **4.2.** Possible impacts of climate on land-use phases

412 Simultaneous broad-scale deforestation phases over large areas caused by land-use may 413 mainly reflect superimposed causes such as climate change (e.g. Berglund, 2003; Tinner et al., 414 2003, 2009). Our study sites showing coeval vegetation succession patterns are both located in landscapes providing fertile arable lands (Guthruf, Zeh, & Guthruf-Seiler, 1999). They are 415 416 mostly lacking major in- and outflows, thus settlements around the lakes, especially at the 417 shores (e.g. pile dwellings) may have been affected by lake level changes (Guthruf, Zeh, & Guthruf-Seiler, 1999). We here assume that people settled at lower elevations preferably near 418 419 lake shores/rivers during warm and dry phases and subsequently moved upwards to slightly higher elevations during cool and wet phases. Thus, the more densely populated warm and 420 421 dry phases might be preferentially reflected in the archaeological and dendrochronological 422 records e.g. at Lake Biel (Fischer, Hafner, Stapfer, Marti, & Affolter, 2017; Hafner & Suter,

2000; Stöckli, 2016; Suter, 2017) and Lake Zurich (Bleicher et al., 2017), whereas the fewer
settlements on dry soils during cool and wet phases (e.g. hilltop sites) were more exposed to
erosion, hence artifacts are scarce.

Common land-use phases across Central and Southern Europe were possibly driven by 426 427 temperature oscillations, which were usually associated with moisture changes. After 8,200 428 cal year BP, warm periods in and around the Alps were generally associated with dry 429 conditions, while decadal to centennial scale cold excursions were wetter, comparable to the 430 conditions during the Little Ice Age (Haas, Richoz, Tinner, & Wick, 1998; Tinner & Lotter, 2006). Indeed wet phases, as reconstructed from lake-level proxies occurred at around 6,300, 5,500 431 and 4,200 cal year BP (Magny, 2004, 2013), when crop production around the lakes was low. 432 433 The linkage to warm phases as reconstructed from tree-ring records (Bircher, 1982, 1986; 434 Renner, 1982) and the Greenland stable oxygen isotope records (Vinther et al., 2006) is less pronounced (Figure 9). Although the Swiss tree-ring records of temperature change are well 435 436 in line with the Greenland stable oxygen isotope record, the chronological precision of these 437 records (c. ±50-100 cal years) is lower than that of the dendrochronologically dated 438 archaeological evidence (c. $\pm 5-10$ years). Previous studies have emphasized the good match 439 between the dendrochronologically dated archaeological phases and solar activity, as reconstructed from the ¹⁴C-residuals at multiannual to multidecadal scales (Gross-Klee & 440 441 Maise, 1997; Kleinmann, Merkt, & Müller, 2015; Magny, 2004, 2013). The link between solar 442 activity and archaeological evidence becomes again evident when comparing the land-use phases at both study sites with the reconstructed average total solar irradiance (TSI; 443 444 Steinhilber, Beer, & Fröhlich, 2009; see Figure 9). Periods with high solar irradiance 445 corresponded with rather dry conditions (low lake levels) and increases of cultural indicators 446 (sum of cereals, L. usitatissimum, P. lanceolata), local-regional fires (macroscopic and

447 microscopic charcoal) and biodiversity (PRI). Conversely, periods with lower TSI and wetter 448 conditions (high lake levels) corresponded to land abandonment phases. We assume that 449 under adverse wet and cool conditions agricultural yields collapsed and human population 450 densities declined over large areas, as shown by the excellent agreement across distant sites 451 (Figure 8). Solar irradiance may have additionally advantaged crop production, given that the 452 cereals, which were the main source of calories, originate from the Near East and are adapted 453 to high insolation.

454

455 **4.3. Implications for global change ecology**

456 Our study shows that over the millennia temperate forests dominated by *F. sylvatica* and *A.* alba were resilient to prehistoric human impacts in the Swiss Plateau. Forest succession after 457 disturbances generally started with light-loving pioneer species (C. avellana, Betula) and was 458 completed within 150-200 years by late successional mixed beech forests. Such 459 palaeoecological information might be crucial to develop sustainable strategies for nature 460 461 conversation and forest management (Feurdean & Willis, 2008; Lindbladh, Niklasson, Karlsson, Björkman, & Churski, 2008; Valsecchi, Carraro, Conedera, & Tinner, 2010), 462 specifically under global-change conditions that are currently releasing forest ecosystem 463 expansions in abandoned rural areas (FOREST EUROPE, 2015). We assume that F. sylvatica 464 forests will remain key communities in Central Europe, if future (summer-) climate warming 465 does not exceed the Holocene variability range (c. +1-2°C compared to the 20th century; Heiri, 466 467 Tinner, & Lotter, 2004). Other trees such as *F. excelsior, Tilia* and *Ulmus* might also re-expand under low disturbance levels (e.g. no browsing, no pollarding) provided that no further 468 469 devastating diebacks due to pathogenic diseases (e.g. Ulmus, F. excelsior) occur (see Brasier, 470 1991; McKinney, Nielsen, Hansen, & Kjær, 2011). Anticipated climate change (Appenzeller et

471 al., 2011; Kovats et al., 2014), however, with generally lower annual precipitation rates may markedly disadvantage F. sylvatica and promote A. alba and Quercus spp. (including 472 submediterranean Q. pubescens) that are less-drought sensitive and thus more competitive 473 under drier conditions (Bugmann et al., 2015; Ellenberg, 2009; Ruosch et al., 2016; Tinner et 474 475 al., 2013). Predicted higher risk of forest fires (Kovats et al., 2014) may affect fire-sensitive A. 476 alba (Tinner et al., 2013), or conversely A. alba and the expansion of mediterranean evergreen 477 trees (e.g. Quercus ilex) may further reduce fire incidence because of positive feedbacks 478 mainly related to decreased biomass flammability (Henne et al., 2015, 2018).

479

480 **5. CONCLUSIONS**

481 We numerically demonstrate for the first time that vegetation dynamics at two distant sites were synchronous over more than 2,000 years, with a pooled sampling resolution of 11 years. 482 483 Interactions between the first agrarian societies and their environment were complex, with humans significantly shaping vegetation structure (creation of open land, shrublands) and 484 composition (shift in species). Conversely, the rise and fall of early farming societies was likely 485 486 dependent on climate. Favourable climatic conditions (i.e. warm and dry summers) probably led to an increase in agricultural yields, the expansion of farming activities and resulting forest 487 openings, whereas unfavourable climatic conditions (i.e. cold and wet summers) likely caused 488 489 crop failures, abandonment of agricultural areas and forest succession. A better understanding of the environmental and societal factors controlling coeveal land-use 490 dynamics as shown in this study would require new climate proxy data (e.g. temperature 491 492 reconstruction from well dated and complete Holocene tree ring series). On the basis of our results and considering the ongoing spread of temperate forests in lowland Central Europe, 493 we conclude that the existing beech forest ecosystems are resilient to anthropogenic 494

disturbances under a changing climate, if the amplitude does not exceed the range of Holocene climate variability. Our results also indicate that abandoned agricultural land can revert to quasi-natural mixed beech forests within 150–200 years. Climate warming exceeding the Holocene variability range, however, has been identified as an important threshold that might cause the rapid collapse of beech forests, giving way to novel communities that are today restricted to Southern Europe.

501

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511

512 AUTHORS' CONTRIBUTIONS

F.R. performed all pollen, macrofossil and quantitative analyses, figure construction, and led
the writing; E.G. and W.T. initiated and designed the research project and wrote substantial
parts of the manuscript; C.S. helped to conduct the biodiversity estimations; A.H. and O.W.
provided archaeological data; W.T. and A.H. obtained funding. All authors contributed
critically to the drafts and gave final approval for publication.

518

519 DATA ACCESSIBILITY

- 520 The data will be available through the Alpine Palynological Database (ALPADABA) after
- 521 publication.

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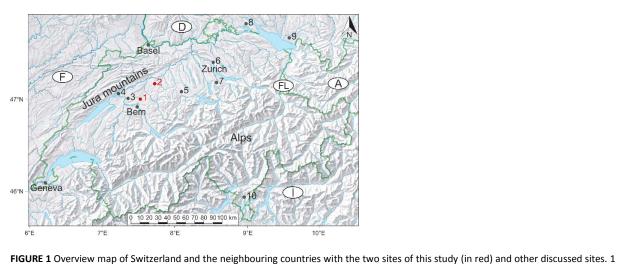
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839 **FIGURES**



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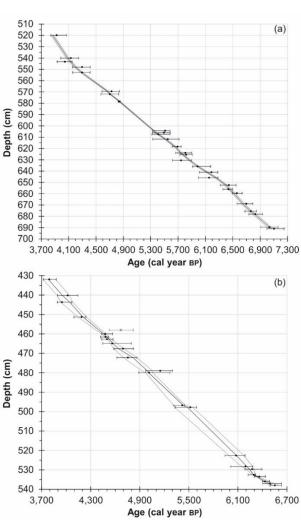
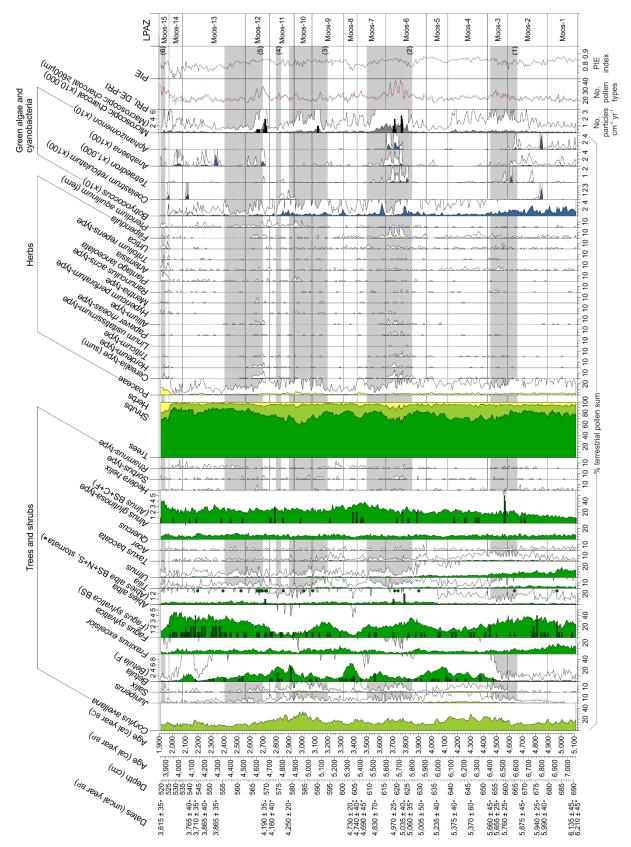




FIGURE 2 Age-depth models of (a) Moossee and (b) Burgäschisee. Black dots show the calibrated ages with 95% (20) probabilities (IntCal13, Reimer et al., 2013). The black lines are the modelled chronologies (OxCal; Bronk Ramsey, 1994, 1995, 2001, 2008; Bronk Ramsey, van der Plicht, & Weninger, 2001; Bronk Ramsey & Lee, 2013). One age at Burgäschisee (dark grey x) was treated as an outlier. The grey lines indicate the 95% (2σ) probabilities of the models.

Moossee; 2 Burgäschisee; 3 Lobsigensee (Ammann, 1989); 4 Lake Biel (Fischer, Hafner, Stapfer, Marti, & Affolter, 2017; Hafner & Suter, 2000; Stöckli, 2016; Suter 2017); 5 Soppensee (Lotter, 1999); 6 Zürich Opéra (Bleicher et al., 2017); 7 Egelsee (Wehrli, Tinner, & Ammann, 2007); 8 Mindelsee (Rösch, Kleinmann, Lechterbeck, & Wick, 2014a, b); 9 Degersee (Kleinmann, Merkt, & Müller, 2015); 10 Lago di Origlio (Tinner, Hubschmid, Wehrli, Ammann, & Conedera, 1999). A = Austria, D = Germany, F = France, FL = Liechtenstein, I = Italy (topographic map: ©

Depth (cm) Depth (cm)



856 857 858 FIGURE 3 Moossee high-resolution sequence. Presented are selected pollen, spore, algae and cyanobacteria percentages, microscopic 859 charcoal influx values, macroscopic charcoal concentrations (concentration = number of particles \geq 600 µm/17.5 cm³, black bars, plotted on 860 top axis), selected plant macrofossil concentrations (concentration = number of remains/17.5 cm³, dark green bars, plotted on top axis), presence of Abies alba stomata (dark green dots), palynological richness (PRI, light blue), evenness-detrended palynological richness (DE-PRI, red) as well as palynological evenness (PIE). Empty curves show 10× exaggerations. Light grey bars indicate the time windows of local archaeological findings reflecting settlement phases (1-6). BS = bud scales, C = cones, F = fruits, N = needles, S = seeds.

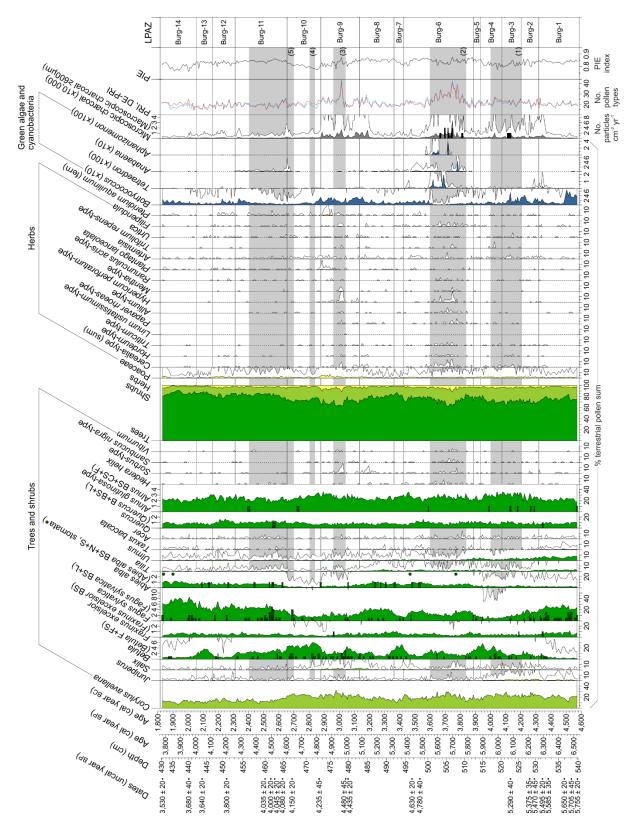




FIGURE 4 Burgäschisee high-resolution sequence. Presented are selected pollen, spore, algae and cyanobacteria percentages, microscopic charcoal influx values, macroscopic charcoal concentrations (concentration = number of particles ≥600 µm/11 cm³ (standard volume), black bars, plotted on top axis), selected plant macrofossil concentrations (concentration = number of remains/11 cm³ (standard volume), dark green bars, plotted on top axis), presence of Abies alba stomata (dark green dots), palynological richness (PRI, light blue), evenness-detrended palynological richness (DE-PRI, red) as well as palynological evenness (PIE). Empty curves show 10× exaggerations. Light grey bars indicate 872 the time windows of local archaeological findings reflecting settlement phases (1–4). B = bud, BS = bud scales, CS = cone scales, F = fruits, FS 873 = fruit scales, L = leaves, N = needles, S = seeds.

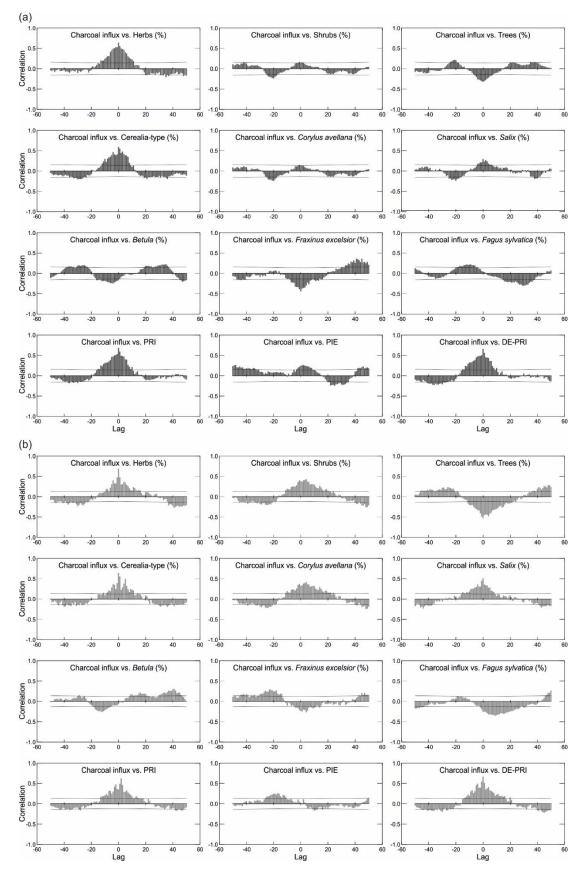
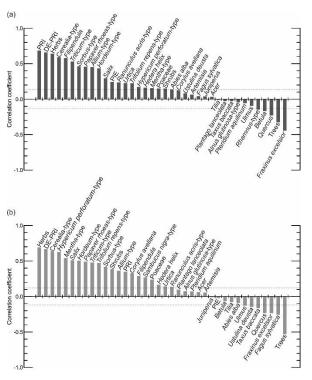
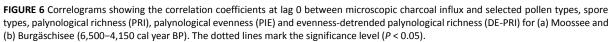




FIGURE 5 Cross-correlation analysis of microscopic charcoal influx vs. percentages of selected pollen types, palynological richness (PRI), palynological evenness (PIE) and evenness-detrended palynological richness (DE-PRI) from (a) Moossee and (b) Burgäschisee (6,500-4,150 cal year BP). 1 lag = 11.1 ± 0.2 cal years (Moossee), 8.7 ± 0.6 cal years (Burgäschisee). The solid black lines mark the significance level (P < 0.05).





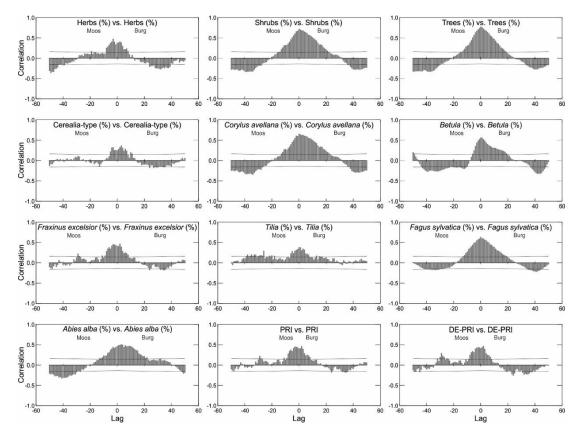


FIGURE 7 Cross-correlation analysis of percentages of selected pollen types, palynological richness (PRI) and evenness-detrended palynological richness (DE-PRI) from Moossee (Moos) vs. percentages of selected pollen types, PRI and DE-PRI from Burgäschisee (Burg) (6,500–4,150 cal year BP). 1 lag = 11.1 ± 0.2 cal. years. The solid black lines mark the significance level (P < 0.05).

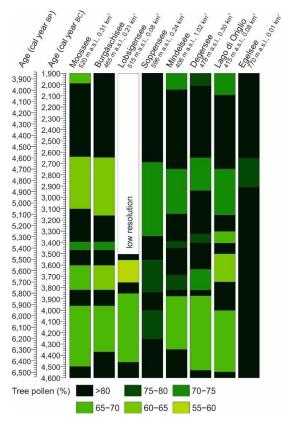
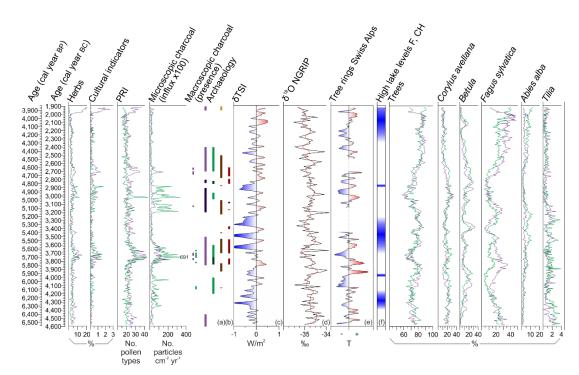


FIGURE 8 Tree pollen percentages from selected sites (see Figure 1). Dark green bars indicate periods with fairly closed forests (tree pollen % >80). Light green bars indicate phases with forest openings (tree pollen % <80).



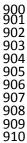


FIGURE 9 Percentages of selected pollen types, palynological richness (PRI), microscopic charcoal influx, macroscopic charcoal presence (\geq 600 µm) and local archaeological findings at Moossee and Burgäschisee compared to palaeoclimatic proxies. Purple graphs = Moossee, green graphs = Burgäschisee. Cultural indicators = Σ Cerealia-type + *Plantago lanceolata* + *Linum usitatissimum*-type. Bottom labelling, center: (a) Local archaeological settlement phases within the city of Zurich (Bleicher et al., 2017). (b) Dendrochronologically dated settlement phases at Lake Biel (Fischer, Hafner, Stapfer, Marti, & Affolter, 2017; Hafner & Suter, 2000; Stöckli, 2016; Suter 2017). Dark coloured bars (archaeology) indicate dendrochronologically dated settlement phases (Moossee, Burgäschisee and Zurich). (c) δ TSI (total solar irradiance) (Steinhilber, Beer, & Fröhlich, 2009). (d) δ ¹⁸O NGRIP (Vinther et al., 2006) with smoothing (loess = 0.05, red line). (e) Calibrated tree ring data from the Swiss Alps (Bircher, 1982, 1986; Renner, 1982). Dotted black lines mark data gaps (f) Periods of high lake levels in Eastern France (F) and Western Switzerland (CH) (Magny, 2004, 2013).