INTRODUCTION

Climate change entails increasing concentrations of greenhouse gases in the atmosphere, rising temperatures, and shifts in precipitation patterns (IPCC, 2014). These changes will profoundly alter individual species’ physiologies and ecology, interaction networks, community composition, and (agro)ecosystem functioning (Abdala-Roberts et al., 2019; Bellard et al., 2012;...
Boukal et al., 2019; Newman et al., 2011; Parmesan & Yohe, 2003; Selvaraj et al., 2013; Walther, 2010). The feeding efficiency of natural enemies of herbivores in particular was predicted to be altered by climate change (Rosenblatt & Schmitz, 2016; Tylianakis et al., 2008; Voigt et al., 2003), which may lead to strong effects on trophic interaction networks and biological control services (Altieri et al., 2012; Bohan et al., 2013; Gravel et al., 2016; van der Putten et al., 2004; Vidal & Murphy, 2018; Woodward & Bohan, 2013). Yet, manipulative studies that investigate the combined impacts of changes in CO2, temperature, and precipitation on members of higher trophic levels remain rare (Jamieson et al., 2012; Laws, 2017; Rosenblatt & Schmitz, 2016), thus limiting our capacity to predict how climate change will affect natural enemies of herbivores.

Natural enemies of herbivores such as predators, parasites, and parasitoids may suffer from climate change via extrinsic and intrinsic mechanisms. Intrinsic mechanisms refer to direct effects of climate-associated abiotic parameters on an organism (Cornelissen, 2011; Robinet & Roques, 2010). Extrinsic mechanisms refer to indirect effects, mediated for instance through shifted distributions, altered phenology or physiology of herbivores and their host plants (Chidawanyika et al., 2019; Damien & Tougeron, 2019; Han et al., 2019; Jeffs & Lewis, 2013; Kaplan et al., 2016; Kharouba et al., 2018; Plincebourde et al., 2017; Renner & Zohner, 2018). While investigating indirect effects of climate change across trophic levels is crucial to predict community shifts (van der Putten et al., 2010), understanding the direct impact of climate change onto each trophic level is essential to characterize the underlying mechanisms (Thomson et al., 2010).

Over the last years, detailed experiments have been conducted to evaluate the direct impact of individual abiotic factors related to climate change on natural enemies of herbivores. For example, temperature was identified as a key determinant of parasitoid developmental rates, survival, fecundity, parasitism, and dispersal (Hance et al., 2007; Selvaraj et al., 2013; Walther et al., 2002). Thermal performance curves revealed that increasing temperatures enhance parasitism success until a maximum at optimum temperature, beyond which any increase in temperature leads to a decline of parasitism success (Chidawanyika et al., 2019; Furlong & Zaluicki, 2017). Optimum temperature was repeatedly reported to be lower for parasitoids than for their hosts, indicating that parasitoids may be more susceptible to elevated temperatures than herbivores, resulting in lowered biocontrol efficacy over time (Furlong & Zaluicki, 2017). Similarly, changes in CO2 levels and precipitation were demonstrated to affect natural enemies of herbivores. While the impact of elevated CO2 on the third level is generally mediated by shifts in bottom-up forces, direct effects of CO2 on predators and parasitoids should not be neglected (Jun Chen et al., 2007). Notably, elevated CO2 alters the chemosensation of natural enemies by disrupting their ability to perceive or process cues from their environment (Draper & Weissburg, 2019), for instance, through a reduction of olfactory receptor sensitivity and habituation of higher-order neurons (Majeed et al., 2014). Changes in precipitation were reported to disturb the physiology and behavior of herbivore natural enemies (Barnett & Facey, 2016; Jamieson et al., 2012; Torode et al., 2016). While arachnids and insects possessing a waxy cuticle can reduce evaporation (Berridge, 2012), soft body organisms (nematodes, isopods, and myriapods) may suffer larger water loss under drought conditions (Sylvain et al., 2014). Drought may also prompt arthropod predators to migrate, hide in the soil, or build shelters to avoid desiccation (Berridge, 2012; Melguizo-Ruiz et al., 2016; Willmer, 1982), thus reducing their foraging activity.

Compared to single climate parameters, much less is known about the direct effects of combined and interactive impact of multiple parameters on natural enemies of herbivores (Hiltrop et al., 2016; Jactel et al., 2019; Kreyling & Beier, 2013; Robinson et al., 2012). A meta-analysis on the impact of multiple global change drivers on the survival of animals from all trophic levels revealed that combined climatic stressors led to nonadditive effects, comprising synergistic and antagonist interactions, in the majority (65%) of the examined studies (Darling & Côté, 2008). The occurrence of synergistic and antagonist effects varies considerably between meta-analyses and is species and sex-dependent (Crain et al., 2008; Darling & Côté, 2008; Rosenblatt & Schmitz, 2014). Studies evaluating the direct effects of combined climatic factors on herbivore natural enemies remain scarce. One example showed antagonist interactions between elevated temperature and decreased precipitation pattern on parasitoid wasp success: While elevated temperature and drought individually enhanced parasitism success, their combination reversed this effect (Romo et al., 2012). Another study, using altitudinal gradient experiments, revealed that warming, elevated nitrogen resources, and their combination resulted in similar increase in total herbivore biomass, but not that of parasitoids (Sassi & Tylianakis, 2012). Even though parasitoids parasitized more herbivores under elevated temperature, this increase was not proportionate to the increase in host abundance. This observation could be explained by the negative effect of both drivers and their subadditive interactions on parasitism rates (Sassi & Tylianakis, 2012). Despite these examples, to date, the outcome of multiple global change factors on the third trophic level remains unpredictable and is likely to lead to "ecological surprises" (Darling & Côté, 2008).

Here, we developed a microcosm climate system to quantify the direct impact of elevated temperature (incl. an appropriate diurnal rhythm) and CO2 as well as reduced precipitation, as predicted by the Representative Concentration Pathway 8.5 (RCP 8.5, (IPCC, 2014)), on the survival and foraging efficiency of different natural enemies of herbivores. Focusing on direct effects comes with the limitation that no preys can be inserted in the system during exposure. Our study therefore investigated the short-term impact of climate change on herbivore enemies. We reveal that exposure to RCP 8.5 conditions has no negative direct effects on several different natural enemies of herbivores, suggesting that the third trophic level may be resilient to direct abiotic effects imposed by climate change.
Habitat

Herbivores. Eggs of the cucumber banded beetle, Diabrotica balteata (Coleoptera: Chrysomelidae), were kindly provided by Oliver Kindler (Syngenta, Stein, Switzerland). Hatching larvae were reared on freshly germinated maize seedlings (var. Akku, DSP, Delley, Switzerland) until use. Third instar larvae were used for all infectivity experiments. Eggs of Egyptian cotton leafworm, Hemithea littoralis (Lepidoptera: Noctuidae), were found in Supplementary Material (Material S1). Falcon tubes (50 ml, Falcon, Greiner Bio-One, Frickenhausen, Germany) were filled with 30 g dry (80°C for 48 hr), sieved (2 cm mesh) soil (40% sand, 35% silt, 25% clay; Landere, Ricoter, Aarberg, Switzerland). The natural soil microbiota was re-implanted to the soil as previously described (Hu et al., 2018). All falcon tubes were placed in dry-bath cyclers (Digital Heating Cooling Drybath, Thermo Scientific, Fisher Scientific AG, Reinach, Switzerland). Ambient air from outside the building was used to mimic current conditions (=450 ppm ± 50 ppm). A concentration of 400 ppm CO₂ (purity 100%, 54.6 L bottle, and pressure of output at 0.8 bars, Gümligen, Switzerland) was added to ambient air (=850 ppm ± 50 ppm) to reach expected RCP 8.5 scenarios. The resulting CO₂: air mix was pushed through a filter of activated carbon (Camozzi, Warwickshire, United Kingdom). The flow rate sent through individual Falcon tubes was adjusted to 1 L/min. The outflow of the Falcon tubes was connected to a collection system, and a CO₂ sensor to verify CO₂ levels. The collected air was then released in the environment.

The temperature in the Falcon tubes was controlled through the dry-bath cyclers and followed a diurnal variation of 3.5°C. Soil temperatures used to mimic current conditions were of 17.8°C at 6 a.m., and gradually increased to reach 21.4°C at 4 p.m. (Figure S1), as reported for the Swiss Plateau over the past two decades (MeteoSwiss, Federal Office of Meteorology and Climatology, Zürich, Switzerland). The temperatures mimicking the RCP 8.5 scenario were set to 21.2°C at 6 a.m. and progressively increased to reach 24.8°C at 4 p.m. (Figure S1).

The moisture present in the tubes was controlled by adding the soil leachates to the tubes once at the beginning of the experiment. The volume of water to add in the tubes was calculated based on the soil density of 1.2 g/cm³. Current moisture levels (23% soil moisture) were achieved by adding 16.6% (v/v) microbiota extracts contained in tap water and 6.4% (v/v) additional tap water. Predicted moisture levels (RCP 8.5, 28% less precipitation, Figure S2) were achieved by adding 16.6% (v/v) microbiota extracts contained in tap water only. The temperatures were adjusted to the different scenarios over a six-hour adaptation period (Figure S1). The water loss over the run of the experiments was not significantly different between treatments.

2.3 | Microcosm systems

To manipulate CO₂ levels, temperature, and moisture, we developed a microcosm system using dry-bath cyclers and a custom-made CO₂ dosage system. A detailed description of the microcosms can be found in Supplementary Material (Material S1). Falcon tubes (50 ml, Falcon, Greiner Bio-One, Frickenhausen, Germany) were filled with 30 g dry (80°C for 48 hr), sieved (2 cm mesh) soil (40% sand, 35% silt, 25% clay; Landere, Ricoter, Aarberg, Switzerland). The natural soil microbiota was re-implanted to the soil as previously described (Hu et al., 2018). All falcon tubes were placed in dry-bath cyclers (Digital Heating Cooling Drybath, Thermo Scientific, Fisher Scientific AG, Reinach, Switzerland). Ambient air from outside the building was used to mimic current conditions (=450 ppm ± 50 ppm). A concentration of 400 ppm CO₂ (purity 100%, 54.6 L bottle, and pressure of output at 0.8 bars, Gümligen, Switzerland) was added to ambient air (=850 ppm ± 50 ppm) to reach expected RCP 8.5 scenarios. The resulting CO₂: air mix was pushed through a filter of activated carbon (Camozzi, Warwickshire, United Kingdom). The flow rate sent through individual Falcon tubes was adjusted to 1 L/min. The outflow of the Falcon tubes was connected to a collection system, and a CO₂ sensor to verify CO₂ levels. The collected air was then released in the environment.

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2.4 | Effects of climate change on aboveground predator and parasite survival

To evaluate the direct impact of current and predicted climatic conditions on success of predators, we evaluated the survival and predation success, defined as foraging efficiency, of a wolf spider (A. albofasciata, n = 25–30), a ladybug (A. bipunctata, n = 17), and a parasitic wasp (A. ervi, n = 5–6) after exposure to climatic conditions in the microcosms. Half of the individuals for each predator species were exposed to current conditions, while the other half was exposed to RCP 8.5 conditions. A piece of cotton soaked in 10% sucrose solution, or a slice of apple was added to the tubes as a sugar source. After three days of incubation, the temperature in all systems ramped to ambient conditions to prevent any temperature-related physiological shock. All herbivore enemies were exposed to ambient temperature levels for one day prior predation assays. This resulted in exposure to current and predicted temperature for three days, and to current or predicted CO₂ and precipitation levels for 4 days prior predation assays. This four-day duration chosen based on the assumption it would be sufficient to trigger any possible impact on the organisms’ physiologies without leading to lethal starvation. All predation assays were performed under ambient conditions. To measure direct effects of climate change only, all herbivores were kept in ambient conditions at all time.

2.5 | Effects of climate change on aboveground predation and parasitism success

To assess the predation and parasitism success of the three aboveground arthropods, four independent assays were conducted as follows. In a first experiment, individual wolf spiders were exposed to current or predicted RCP 8.5 climatic conditions as described above. After exposure, each spider was placed in a solo cup (250 ml, Pack Markt Sabaratnam, Solothurn, Switzerland). Half of the cups contained ten second instar S. littoralis caterpillar larvae, while the second half contained ten D. melanogaster flies (n = 9–12). In a second assay, individual ladybugs were exposed to current or RCP 8.5 climatic conditions (n = 9–14) and then placed in Petri dishes (94 x 16 mm, Bio-One Petri Greiner, Huberlab, Switzerland) containing a piece of cotton soaked in a 10% sucrose solution and thirty aphids R. padi. In the third experiment, three ladybugs per tube were exposed to current or predicted RCP 8.5 conditions (n = 5–6) and then placed in Petri dishes containing a piece of cotton soaked in a 10% sucrose solution and thirty aphids B. brassicae. In these assays, the number of consumed preys was recorded 24 hr after introduction of the predator. In a fourth assay, nine to twelve parasitoid wasps were placed in each tube and exposed to either current or predicted RCP 8.5 conditions (n = 5–6). The wasps were then placed into Petri dishes containing a slice of apple and thirty aphids B. brassicae. The wasps and aphids were left together for five days before assessing the aphid parasitism rate.

2.6 | Climate change on belowground parasitism rates

To investigate direct effects of an extreme climate change scenario (RCP 8.5) on the ability of soil-dwelling parasites to infect an herbivore, five strains of the entomopathogenic nematode H. bacteriophora (strains EN01, TT01, DE6, HU2, PT1) were tested in the microcosms described above. Briefly, 3,000 infective juvenile nematodes, suspended in one mL tap water, were added in each microcosm, and exposed to current or RCP 8.5 climatic conditions for three days. After a one-day incubation period to ambient temperature levels, the soil from the microcosms was collected and poured into a plastic cup (250 ml). The soil moisture was then adjusted to 23% in all treatments. Ten third instar larvae of the root herbivore D. balleata were added into each plastic cup. After seven days, all larvae were placed in white traps (White, 1927). The parasitism rate and number of emerging juveniles were recorded after 14 days. The number of EPNs emerging from larvae infected with EPNs originating from one tube was averaged and used as a proxy for EPN fitness. Testing five strains with sufficient replicate numbers required four independent experiments, each including EN01 and two additional strains (n = 7–16 per strain and treatment in total). The infectivity and number of EPNs emerging from the insect cadavers were expressed relatively to EN01-ambient conditions treatment.

2.7 | Effects of single climatic conditions on EPN performance and fitness

To appraise the individual and interactive effects of soil temperature, soil moisture, and CO₂ onto EPN performance and fitness, a full factorial experiment was designed, using all combinations between current and predicted climatic variable levels. The assays were performed using the nematode H. bacteriophora strain EN01 and following the same methods as described above. In a first assay, EPN survival was recorded by collecting EPNs using a modified Baermann funnel method (Baermann, 1917). Briefly, soil samples were placed into a paper towel and rehydrated until saturation in a funnel, which was sealed with metal clips at the bottom. EPN extraction was performed at ambient temperature conditions. Extracts with suspended EPNs were collected after 24 and 48 hr. The proportion of EPNs collected after 24 hr over the total recovered EPN number was used as a proxy for mobility. In a second assay, EPNs were collected and used for infectivity and fitness assessment as described above. This experiment was repeated three times to ensure good replicate numbers (n = 6–12 per treatment).

2.8 | Statistical analyses

Statistical analyses were conducted using R (version 3.5.3, https://www.r-project.org) and online tools (http://quantpsy.org; https://www.graphpad.com). Normality and heteroscedasticity of error variance were assessed using Levene’s and Shapiro–Wilks tests, as
well as by visualizing quantile–quantile plots and model residuals versus fitted values. The survival rates of spiders and ladybugs were analyzed using chi-squared tests incorporating a Yates correction for continuity. The survival rate of wasps was analyzed using a generalized linear model using the proportion of surviving wasps in each tube as a response to climatic scenarios. Spider predation rates were analyzed using linear models. Ladybug predation and wasp parasitism rates were analyzed using generalized linear models with a binomial error structure. EPN infectivity and emergence data were expressed relatively to the EN01 strain in ambient conditions. EPN emergence data from insect infected with EPNs originating from the same microcosm tube were averaged and ln transformed prior analysis. One emergence data point was considered as outlier (Grubbs’ test, \( p < .05 \)) and removed from the analysis. EPN relative infectivity and fitness were analyzed using linear models. Climatic scenarios/variables, herbivore species (when compared within a single experiment), and nematode strains were treated as fixed effects. Cyclers and CO\(_2\) channels did not increase the model gain on likelihood and were therefore removed from the models. Experimental repetitions were treated as fixed variables. Significant models were evaluated by Tukey Honest Significant Differences (TukeyHSD) post hoc tests.

3 | RESULTS

3.1 | Predicted climatic conditions do not directly influence the survival of aboveground predators

The designed microcosms ensured good survival rates of wolf spiders, ladybugs, and parasitoid wasps over the course of the experiment (Figure 1). Exposure to RCP 8.5 climate conditions, including elevated levels of CO\(_2\), elevated temperature, and decreased moisture, did not impair the survival of any of the tested arthropods (Figure 1a, c, e).

![Figure 1](image-url) Survival and predation efficacy of herbivore natural enemies after exposure to current and predicted RCP 8.5 climate scenario (IPCC, 2014). Survival (Yates’ Chi square, mean ± se, \( n = 20–24 \)) and consumption rate (linear models, means ± se, \( n = 9–12 \)) of the wolf spider, Alopecosa albofasciata, after exposure to temperature, precipitation and CO\(_2\) levels of current or predicted (RCP 8.5) climate scenarios (a, b). Survival (Yates’ Chi-square, means ± se, \( n = 17 \)) and consumption rate (generalized linear models, means ± se, \( n = 6–14 \)) of the ladybird, Adalia bipunctata, after exposure to temperature, precipitation, and CO\(_2\) levels of current or predicted (RCP 8.5) climate scenarios (c, d). Survival (generalized linear model, means ± se, \( n = 5–6 \)) and infection rate (generalized linear models, means ± se, \( n = 5–6 \)) of the parasite wasp, Aphidius ervi, after exposure to temperature, precipitations, and CO\(_2\) levels of current or predicted (RCP 8.5) climate scenarios (e, f). Survival rate (generalized linear model, means ± se, \( n = 2–3 \)) and infection rate (generalized linear models, means ± se, \( n = 6–9 \)) of the commercial entomopathogenic nematode, strain EN01, under exposure to temperature, moistures and CO\(_2\) corresponding to current or predicted (RCP 8.5) climate scenarios (g, h). Stars indicate significant differences, *: \( p < .05 \). Dots indicate trends:. 0.05 < \( p < .10 \).
3.2 | Predicted climatic conditions altered aboveground predation in a species-specific manner

Exposure to current or predicted RCP 8.5 climatic conditions resulted in species-specific changes in predation success. The combination of elevated CO₂, increased temperature, and decreased moisture (RCP 8.5) altered the spider predation rates in a prey-specific manner. It slightly increased spider consumption of the cotton leafworm, *S. littoralis*, but not of the fly larvae, *D. melanogaster* (Figure 1b). Exposure of ladybugs to current or predicted RCP 8.5 climatic conditions resulted in similar consumption rates of cabbage- and bird cherry-oat aphids (Figure 1d). Similarly, exposure to current or predicted RCP 8.5 climatic conditions did not alter the wasp parasitism rate of cabbage aphids (Figure 1f).

3.3 | Predicted climatic conditions increase herbivore infection rates by nematodes

Direct exposure to current or predicted RCP 8.5 climatic conditions in microcosms did not decrease infection rates across five tested EPN strains (Figure 2a). On average, exposure to RCP 8.5 conditions increased EPN infection rates by 10.03%. The effect of the RCP 8.5 scenario was the strongest on EN01, whose infection rates increased by 21.44% compared to current conditions. Exposure to predicted RCP 8.5 scenarios did not alter the number of emerging juveniles per insect host in any of the tested strains, nor the overall progeny number (Figure 2b-c).

3.4 | Reduced soil moisture decreases EPN mobility, but increases EPN fitness

Factorial manipulation of the individual climatic variables revealed that exposure to lower soil moisture levels (~28% precipitation) modulated EPN success, while exposure to increased temperature and elevated CO₂ levels had no additional effects. Reduced soil moisture slightly decreased EPN survival (albeit not significantly), and mobility (Figure 3a-b), but increased infection rates (Figure 3c). While the impact of soil moisture on the progeny number per host was negligible (Figure 3d), the overall EPN fitness was increased (Figure 3e).

4 | DISCUSSION

Our study highlights that a short exposure to temperature, CO₂ and precipitation levels predicted under an extreme climatic scenario (RCP 8.5) does not directly impair the efficacy of different natural enemies of herbivores, including several important biological control agents. We even observed a prey species-specific increase in spider...
efficiency and enhanced nematode infection rates. This finding suggests that the negative impact of climate change on higher trophic levels observed in large meta-analyses and field experiments may mostly be mediated through either long-term exposure and/or indirect (plant and herbivore-mediated) changes. The possible underlying mechanisms and ecological relevance of these observations are discussed below.

Current models predict that climate change will strongly harm higher trophic levels (Thakur, 2020; Voigt et al., 2003). Our study reveals that short-term climatic-related changes have negligible direct effects on the ability of third trophic level organisms to predate or parasitize herbivores. The negative impact of climate change on higher trophic levels observed in meta-analyses and field experiments may therefore be mediated through either long-term exposure or indirect changes. Exposure to future climatic scenario did not affect the survival of aboveground herbivore natural enemies such as spiders, ladybugs, and parasitoid wasps but decreased the survival of belowground entomopathogenic nematodes (EPNs) within four days. This observation concurs with the hypothesis that arthropods, possessing a waxy cuticle that protects them from desiccation, may be more resistant to climate change than soft-bodied organisms (Sylvain et al., 2014). More investigation would be required to confirm this observation. While arthropod predators and parasites are prevalent aboveground (Nyffeler & Birkhofer, 2017), EPNs are key herbivore enemies belowground (Denno et al., 2008; Půza & Mrácek, 2005). The contrasting impact of climate change on above- and belowground organisms survival supports the hypothesis that green, aboveground, food webs may be more sensitive to climate change than brown, belowground ones (Thakur, 2020). Yet, future climatic conditions did not negatively impact the predation or parasitism rates by herbivore natural enemies, and even increased the success of nematodes and spiders. The increased infection rate by nematodes was sufficient to compensate for their lower survival in term of progeny number. While climate change may not directly impair EPN ability to control herbivore pest population, changes in host abundance, phenology and quality can drastically modulate their efficacy (Guyer et al., 2018; Hiltpold et al., 2016, 2020; Půza & Mrácek, 2005). Interestingly, the increase in spider predation rate
was prey species-specific and may reflect changes in prey preferences. Such shift in prey consumption was previously observed in Arctic spiders, which ultimately led to slower decomposition rates in soil (Koltz et al., 2018). As spiders consume up to 800 million tons of prey per year worldwide (Nyffeler & Birkhofer, 2017), identifying future shifts in their predation strategy will be crucial to reliably predict changes in food webs and ecosystem functioning.

Understanding how different climatic variables act together to impact the efficiency of natural enemies of herbivores is an important frontier in multitrophic interaction ecology. Investigating the impact of single and combined climate-related drivers onto EPN survival and efficacy revealed no interaction between elevated temperature, CO₂, and decreased precipitation. While temperature is known to be a major driver of EPN survival and infectivity (Aatif et al., 2020; Lalramliana & Yadav, 2016; Pervez et al., 2015), we did not find any impact of warming in our study. This discrepancy may be explained by the low range of temperature used in our study (19.6–23°C) compared to the described beneficial temperature range (25–30°C) (Pervez et al., 2015). Consistently with previous literature, elevated levels of CO₂ did not affect EPN survival and parasitism (Hiltplold et al., 2020). Instead, lower precipitation alone seemed to contribute to increased infection rates and progeny number, while decreasing the nematode survival and mobility in the soil. Nematodes have evolved physiological and behavioral strategies to cope with low environmental moisture (Grewal et al., 2011; Kagimu et al., 2017). They can respond to desiccation by initiating anhydrobiosis, a phenomenon largely associated with the accumulation of trehalose and water stress-related proteins (Grewal et al., 2011; Kagimu et al., 2017; Womersley, 1990). The tolerance to desiccation conferred by anhydrobiosis varies considerably between EPN species and strains (Grewal et al., 2002; Mukuka et al., 2008). Furthermore, EPNs can avoid desiccation through aggregative movement patterns (Ruan et al., 2018) and migration to deeper soil layers where moisture levels are higher (Salame & Glazer, 2015). It was also suggested that EPNs can avoid desiccation by remaining longer inside their hosts (Brown & Gaugler, 1997; Koppenhöfer et al., 1997; Půza & Mráček, 2005, 2007). It is tempting to speculate that the higher infection rates and progeny numbers observed in our study upon predicted climatic scenario are the result of some EPN desiccation tolerance strategy.

Climate change is a multifactorial phenomenon that may affect multitrophic interactions through direct and indirect effects. Interactive effects of the global change drivers and cascading effects among food webs render predictions about the response of multitrophic challenging to draw. Our study complements current models and shows that combined elevated temperature, CO₂ levels and decreased precipitation do not have a direct negative impact on the performance of third trophic level organisms upon short-term exposure. Together with the current literature, this suggest that indirect, bottom-up-mediated, climate change effects may be stronger than direct effects on predator and parasite efficacy in controlling herbivore populations. Yet, our study highlighted possible shifts in prey species, while the latter were kept in ambient conditions. This phenomenon may have substantial repercussions on ecosystem functioning and would deserve more attention. Building reliable models of food web response to climate change will be of crucial importance to face future natural and agricultural challenges.

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CONFLICT OF INTEREST
The authors declare having no competing interests.

AUTHOR CONTRIBUTIONS
Cong van Doan: Data curation (lead); Formal analysis (equal); Investigation (lead); Methodology (supporting); Visualization (equal); Writing-original draft (lead); Writing-review & editing (supporting). Marc Pfander: Methodology (equal); Software (lead); Writing-original draft (supporting). Anouk Sabine Guyer: Conceptualization (supporting); Data curation (supporting); Investigation (supporting). Xi Zhang: Data curation (supporting); Investigation (supporting); Writing-review & editing (supporting). Corina Maurer: Data curation (supporting); Investigation (supporting); Resources (supporting). Christelle A.M. Robert: Conceptualization (lead); Formal analysis (equal); Methodology (equal); Project administration (lead); Resources (lead); Supervision (lead); Visualization (equal); Writing-original draft (supporting); Writing-review & editing (lead).

DATA AVAILABILITY STATEMENT
All raw data are available in the supplementary material (Data S1).

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SUPPORTING INFORMATION
Additional supporting information may be found online in the Supporting Information section.

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